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4 **Comparison of Standardized Sampling and Measurement**  
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7 **Reference Systems for Aircraft Engine**  
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9 **Non-volatile Particulate Matter Emissions**  
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9 **Abstract**

10 The International Civil Aviation Organization has established new regulatory standards for  
11 emissions certification of non-volatile particulate matter (nvPM) from aircraft turbine engines. The  
12 adoption of the nvPM emissions regulatory standards required development of a standardized  
13 sampling and measurement methodology, and rigorous testing. Three reference systems for aircraft  
14 engine nvPM emissions measurement, compliant with the specifications for the standardized  
15 methodology, were independently developed. This paper reports the results of the first inter-  
16 comparison of these three reference systems using a CFM56-7B26/3 aircraft engine to establish  
17 repeatability and intermediate precision of the sampling and measurement systems as part of the  
18 multi-agency international collaborative projects: Aviation-Particle Regulatory Instrumentation  
19 Demonstration Experiment (A-PRIDE) 5/ Studying, sAmpling and Measuring of aircraft  
20 ParticuLate Emissions (SAMPLE) III - SC03. The instruments used in the three reference systems  
21 recorded nvPM mass and number concentration, which were converted to their respective emission  
22 indices for comparison. The reference systems generally agreed to within 15% of the average  
23 nvPM number emission index and 30% of the average nvPM mass emission index. The only  
24 exception was for the nvPM mass instruments, which exhibited a higher variation as the  
25 concentration levels approached the limit of detection. The additional measured particle size  
26 distributions could be approximated to lognormal distributions with the geometric mean diameter  
27 ranging from 15 nm to 38 nm, and the geometric standard deviation varying between 1.53 and  
28 1.92. The results from this study are a benchmark for the variability in standardized sampling and  
29 measurement systems for measuring aircraft engine nvPM emissions.  
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50 **Highlights**

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- 53 • Comparison of three reference systems for measuring aircraft engine nvPM emissions
  - 54 • The nvPM number emission index was generally within 15% of the average value
  - 55 • The nvPM mass emission index was generally within 30% of the average value
  - 56 • Mass-based emissions exhibited high variability towards the instrument limit of detection
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4 **Keywords**

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6 Aircraft engines, non-volatile particulate matter, aviation emissions, black carbon, particle  
7 number, size distributions  
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11 **1 Introduction**

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13 Aircraft engine gaseous and particulate matter (PM) emissions are a unique source of  
14 pollution compared to other sources in the urban environment. The public awareness about  
15 aviation emissions has grown since the rapid increase in demand for commercial air travel in the  
16 1960s, which led to the introduction of emission standards by the International Civil Aviation  
17 Organization (ICAO). Prior to 2010, the ICAO Committee on Aviation Environmental Protection  
18 (CAEP) developed emission standards and recommended practices (SARPs) limited to the  
19 emissions of gaseous pollutants such as oxides of nitrogen (NO<sub>x</sub>), carbon monoxide (CO), and  
20 unburnt hydrocarbons (UHC), and emissions of smoke (reported in terms of smoke number, SN).  
21 The SARPs are applicable to turbojet and turbofan engines with maximum sea-level static rated  
22 thrust >26.7 kN for gaseous emissions and to all engine sizes for smoke emissions. The SARPs,  
23 intended to mitigate the impact of aircraft engine emissions on local air quality, have been  
24 established for the type certification of aircraft engines and are documented with approved test and  
25 measurement procedures in ICAO Annex 16 to the Convention on International Civil Aviation,  
26 Volume II (ICAO, 2017). The type certification process involves operating one or multiple  
27 representative engines of a specific model on a test stand at combustor inlet temperatures  
28 corresponding to the four thrust settings of the standardized Landing and Take-off (LTO) cycle –  
29 7% (taxi), 30% (approach), 85% (climb), and 100% (take-off) – corrected to International Standard  
30 Atmosphere (ISA) conditions (ICAO, 1993). The engine manufacturers submit emissions data,  
31 acquired during these engine type certification tests to the certifying authority for approval and  
32 subsequently for inclusion in the ICAO Aircraft Engine Emissions Databank (EEDB) maintained  
33 by the European Union Aviation Safety Agency (EASA) (EASA, 2019a).  
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51 Since aircraft engine emissions regulatory standards were first adopted in 1981,  
52 ICAO/CAEP has developed increasingly stringent standards for NO<sub>x</sub> emissions. This, along with  
53 the introduction of newer, more fuel-efficient engine technologies, has resulted in lower aircraft  
54 engine emissions over time (Wey & Lee, 2018). However, with the growth of commercial and  
55 cargo air traffic at a rate of ~4.5% per year (Airbus, 2018; Boeing, 2018; IATA, 2019) the absolute  
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4 emissions from the aviation sector have increased, including PM emissions which are forecast to  
5 increase over the next twenty years if current engine technology continues to be used (EASA,  
6 2019b). This in turn has led to several studies to measure, model, and ultimately develop solutions  
7 to mitigate aviation-related emissions (Masiol and Harrison, 2014).  
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11 The smoke standard was introduced at a time when the visibility of exhaust plumes from  
12 aircraft engines was of significant concern for airport safety. Smoke number is determined by  
13 measuring the reduction in reflectance of a Whatman #4 filter used to sample a prescribed mass of  
14 exhaust per unit area of the filter, i.e. 16.2 kg of exhaust gas/m<sup>2</sup>. The precision of the SN  
15 measurement method is reported as ±3 SN (SAE, 2011), which in some cases is higher than the  
16 recorded SN for modern turbofan engines. Also, the SN does not provide any information about  
17 the number, size, and composition of the PM emissions that are required for environmental impact  
18 assessments. In the absence of fleet-wide aircraft engine PM emissions data, the First Order  
19 Approximation (FOA) methodology was developed based on correlations between SN reported in  
20 the ICAO EEDB and available data for non-volatile particulate matter (nvPM) mass emissions  
21 (Wayson, Fleming, & Iovinelli, 2009). FOA and subsequent updates (current version FOA 4.0)  
22 have been used to estimate PM emissions from certified commercial aircraft engines during the  
23 LTO cycle at airports (Rissman et al., 2013; Winther et al., 2015; Woody et al., 2016).  
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27 The eighth meeting of CAEP (ICAO, 2010) recognized the need for PM emissions  
28 regulation in addition to the existing set of regulated pollutants. The first new regulatory standard  
29 for aircraft engine nvPM emissions was adopted by CAEP in 2016 (CAEP/10). It included  
30 reporting requirements and a nvPM mass emissions standard for in-production aircraft engines  
31 with rated thrust >26.7 kN on or after 1 January 2020. The CAEP/10 standard was set at a  
32 regulatory level that matched the existing smoke number visibility standard. The regulatory level  
33 for the CAEP/10 maximum nvPM mass concentration was developed based on a statistical  
34 relationship between nvPM mass concentration and smoke number (Agarwal et al., 2019). This  
35 ensured that any aircraft engine that would have met the certification requirements for smoke  
36 number would also meet the nvPM mass emissions level (all-pass standard). From 1 January 2023,  
37 regulatory limits for mass and number-based nvPM emissions will become applicable for in-  
38 production and new commercial aircraft engine types with a maximum rated thrust > 26.7 kN  
39 (CAEP/11 standard) (ICAO, 2019). Along with this new standard and since the CAEP/10 standard  
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4 includes control of exhaust plumes visibility, CAEP agreed that the smoke number standard will  
5 no longer be applicable for these engines from 1 January 2023.  
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8 Before regulatory standards for nvPM mass and number emissions could be defined, a  
9 standardized sampling and measurement methodology was required for aircraft engine emissions  
10 certification tests. The Society of Automotive Engineers (SAE) Aircraft Engine Gas and  
11 Particulate Emissions Measurement Committee (E-31) was tasked with recommending a  
12 standardized protocol for aircraft engine nvPM mass- and number-based emissions. The  
13 development of this standardized protocol for aircraft engine nvPM emissions was the culmination  
14 of several years of effort to investigate sampling methods, evaluate measurement technologies,  
15 and assess engine type and fuel composition differences on nvPM emissions during the  
16 PARTEMIS (Petzold et al., 2003), APEX (Lobo et al., 2007; Timko et al., 2010; Kinsey et al.,  
17 2010), AAFEX (Kinsey et al., 2012), SAMPLE (Petzold et al., 2011; Crayford et al., 2012;  
18 Crayford et al., 2013; Boies et al., 2015), A-PRIDE (Durdina et al., 2014; Lobo et al., 2015a; Brem  
19 et al., 2015), and MERMOSE (Delhaye et al., 2017) projects, in addition to other studies (Lobo et  
20 al., 2015b). The results informed the development and publication of Aerospace Information  
21 Report (AIR) 6241 (SAE, 2013) and Aerospace Recommended Practice (ARP) 6320 (SAE, 2018).  
22 The standardized protocol is limited to only nvPM emissions, defined as particles that exist at the  
23 aircraft engine exhaust nozzle exit plane that do not volatilize at temperatures greater than 350°C  
24 (ICAO, 2017). Controlling aircraft engine nvPM emissions at the source will lead to lower  
25 emissions on local, regional, and global scales. Total PM emissions downstream of aircraft engines  
26 are not currently considered since they have been shown to vary in space and time as the exhaust  
27 plume cools and expands (Lobo et al., 2012; Timko et al., 2013; Beyersdorf et al., 2014), making  
28 it complicated to develop a standardized sampling and measurement protocol that reports  
29 repeatable concentrations, and to enforce through a certification requirement. The approach to  
30 limit the sampling and measurement system to nvPM was similar to that adopted by the Particle  
31 Measurement Programme (PMP) for the regulation of the number concentration of solid (non-  
32 volatile) particles with a diameter >23 nm emitted from automotive engines (Giechaskiel et al.,  
33 2012).  
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55 Analogous to the reference “golden” particle number measurement system used in PMP  
56 (Martini, Giechaskiel, & Dilara, 2009), three reference systems for aircraft engine nvPM emissions  
57 measurement were independently developed - the Swiss (CHE) fixed (later mobile) system by  
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4 Empa, the European (EUR) mobile system by Cardiff University, and the North American (NAM)  
5 mobile system by Missouri University of Science and Technology. All three reference systems  
6 were compliant with the specifications for the standardized system detailed in AIR6241, but not  
7 identical since AIR6241 permits tolerances for different components and specification ranges. It  
8 was essential to inter-compare the three reference systems using a common aircraft engine source  
9 to establish repeatability and intermediate precision of the sampling and measurement systems,  
10 and to estimate some of the uncertainties associated with the measurements of aircraft engine  
11 nvPM emissions.  
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19 In this paper, we present results of the Aviation-Particle Regulatory Instrumentation  
20 Demonstration Experiment (A-PRIDE) 5/ Studying, sAmpling and Measuring of aircraft  
21 ParticuLate Emissions (SAMPLE) III - SC03 campaign, the first multi-agency international  
22 collaborative project to inter-compare and evaluate the robustness of the Swiss, European, and  
23 North American standardized reference systems for the sampling and measurement of aircraft  
24 engine nvPM emissions. The measurements were performed from 28 July to 25 August 2013, on  
25 a leased CFM56-7B26/3 engine used during dedicated engine testing at the SR Technics engine  
26 test facility in Zürich, Switzerland.  
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## 33 34 35 **2 Methods**

### 36 37 *2.1 CFM56-7B26/3 engine*

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39 The engine used for the dedicated tests was a CFM56-7B26/3, which was leased for the  
40 duration of the campaign. The CFM56-7B engine is the sole powerplant for the Boeing 737NG  
41 family, and it is the most widely used engine in commercial aviation. The “/3” configuration  
42 (improved durability and emissions) was emissions-certified for gaseous pollutants and smoke in  
43 2006. The specific engine used during this campaign had accumulated 5009 flight hours during  
44 2000 flight cycles. The engine had representative operating and performance characteristics with  
45 minimal degradation and negligible oil consumption in the range of a brand-new engine. This  
46 engine was declared to be a representative reference engine to be used as a source of nvPM to  
47 inter-compare the standardized reference systems. Aircraft engine emissions certification-like tests  
48 for nvPM emissions were also performed with a single system (Durdina et al., 2017).  
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## 2.2 *Fuel*

The Jet A-1 fuel used for the emissions tests was provided by SR Technics. Seven fuel samples were collected during the measurement campaign, and the results of the fuel analysis performed by Intertek AG (Schlieren, Switzerland) for selected properties are presented in Table 1. The properties of the Jet A-1 fuel were all within the allowable range specified by ASTM D1655 (ASTM, 2019). The fuel also met all the specifications for emissions certification tests (ICAO, 2008) except that for naphthalenes content, which was below the lower limit requirement at the time (1%). This was also the case for a previous campaign conducted at SR Technics (Lobo et al., 2015a). The lower limit for naphthalenes specification was subsequently changed to 0% (ICAO, 2017). Overall, the different batches of Jet A-1 fuel used for the emissions tests had similar properties, thus eliminating fuel composition as a variable in the calculation of emission indices (EIs) and inter-comparison of the reference systems. An average hydrogen to carbon (H/C) ratio of 1.95 was used for the calculation of the number and mass emission indices for all three reference systems.

**Table 1: Properties of Jet A-1 fuel used during the A-PRIDE 5/SAMPLE III-SC03 campaign**

Property	Units	Method	Allowable Range	Test Fuel Samples <sup>†</sup>						
				29 Jul 2013	2 Aug 2013	5 Aug 2013	12 Aug 2013	17 Aug 2013	18 Aug 2013	25 Aug 2013
Density at 15°C	kg/m <sup>3</sup>	ASTM D4052	780-820	797.6	797.6	797.8	797.8	797.8	797.8	797.2
Kinematic viscosity at -20°C	mm <sup>2</sup> /s	ASTM D445	2.5-6.5	3.591	3.618	3.598	3.596	3.599	3.599	3.618
Distillation temperature	°C	ASTM D86	155-201	169	168	168	168	167	168	169
10% boiling point			235-285	265	265	265	261	263	264	264
Final boiling point										
Net heat of combustion	MJ/kg	ASTM D3338	42.86-43.50	43.3	43.3	43.3	43.3	43.3	43.3	43.3
Aromatics	volume %	ASTM D1319	15-23	17.7	17.7	17.4	18.0	17.7	17.7	17.5
Naphthalenes	volume %	ASTM D1840	0-3% <sup>‡</sup>	0.68	0.72	0.71	0.71	0.74	0.75	0.70

<sup>†</sup> All fuel samples were collected from the fuel line in the test cell except the on 5 Aug 13, which was collected directly from the fuel tanker

<sup>‡</sup> Original allowable range 1.0-3.5%, subsequently updated to 0-3%



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Smoke point	mm	ASTM D1322	20-28	21	21	21	21	22	21	22
Hydrogen content	mass %	ASTM D5291	13.4-14.3	14.18	14.18	14.28	14.04	13.96	14.00	13.76
Sulphur content	mass %	ASTM D5453	< 0.3%	0.053	0.033	0.039	0.039	0.042	0.042	0.042
H/C ratio (calculated)			1.84-1.99	1.97	1.97	1.99	1.95	1.93	1.94	1.90

## 2.4 Experimental Setup

A schematic of the experimental setup is shown in Fig. 1. Details of the various sections of the sampling and measurement systems are provided in the following sections.

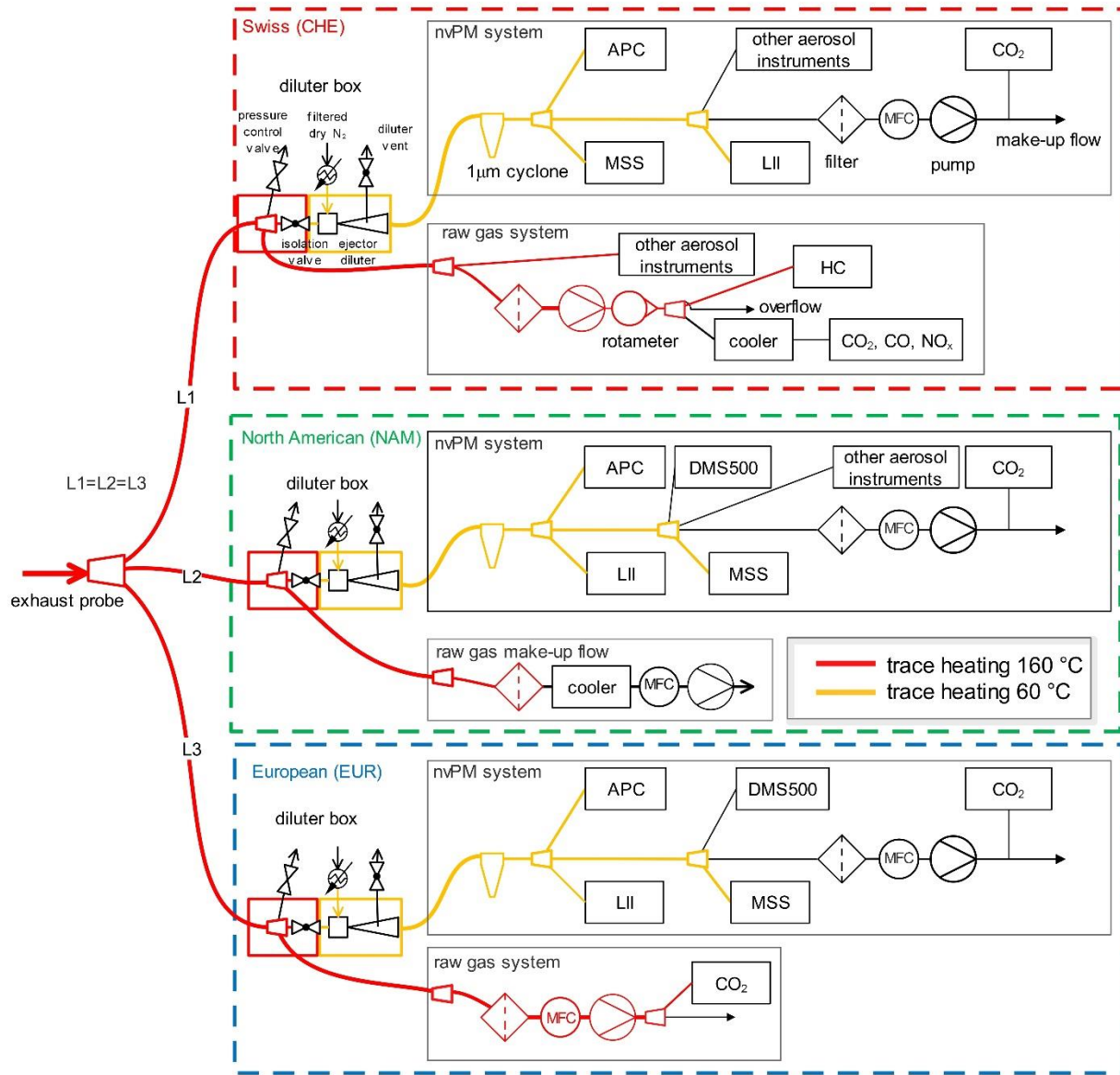


Figure 1: Experimental setup for CFM56-7B26/3 engine tests (not to scale)

#### 2.4.1 *Extractive sampling probes*

Two exhaust sampling probe assemblies were used to extract emission samples within 1 m of the engine exit plane. A fixed multi-point probe made of Inconel 625 alloy compliant with the requirements for nvPM sampling and measurements for emissions certification (ICAO, 2017) was used during a subset of the dedicated engine tests. It consisted of a cruciform with six sampling orifices located on each of the four arms, from which three orifices per arm were used to comply with the carbon balance check (air/fuel ratio estimated from the exhaust sample total carbon concentration agrees with the estimate based on engine air/fuel ratio). The sampling orifices were symmetrically located on circular radii from the center of the cruciform. Various configurations of the number and position of sampling orifices on the multi-point probe were evaluated during the campaign. The primary goal of this assessment was to ensure that the three reference systems operating in parallel were adequately supplied with sample flow, while still providing an exhaust representative sample and complying with the carbon balance check.

A traversable single-point probe used during previous campaigns at SR Technics (Lobo et al., 2015a) was used during piggy-back measurements and a subset of dedicated engine tests. The probe with an 8 mm ID orifice was made of Inconel 600 alloy. The probe's vertical traverse capability afforded representative sampling at the exit plane of different engine types operated in the test cell (for the CFM56-7B26/3 engine, the distance from the probe to the exit plane was 0.8 m). The specific vertical sampling location was optimized during testing, ensuring the carbon balance of the single-point probe maintained satisfactory agreement (< 10%) with the multi-point probe at all similar test conditions.

#### 2.4.2 *Sampling and measurement systems*

The probe was connected to a three-way splitter using a 5.2 m long, 8 mm ID thin-walled stainless steel tubing, electrically trace heated and insulated to maintain a temperature of  $160^{\circ}\text{C} \pm 15^{\circ}\text{C}$ . The exhaust samples were distributed to the three reference systems. Each reference system, while compliant with the specifications detailed in AIR6241, was independently assembled from non-identical components and evaluated prior to deployment in the campaign. Briefly, each system had a three-way splitter to distribute the exhaust sample to the line for raw gaseous emission measurements, the pressure control line to regulate the sample inlet pressure, and the nvPM line for nvPM mass- and number-based emission measurements (Fig. 1). Of the three reference

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4 systems, only the NAM system did not have measurements of raw gaseous emissions on the raw  
5 exhaust line. Instead, a pump was used to draw flow down the gaseous emissions line in order to  
6 achieve a prescribed flow velocity in each leg of the splitter, and to be consistent with the CHE  
7 and EUR system flowrates. The undiluted CO<sub>2</sub> concentration data from the CHE reference system  
8 were used to calculate dilution factors for all three systems. The variability in undiluted CO<sub>2</sub>  
9 concentrations measured by the CHE and the EUR reference systems was evaluated. The slope of  
10 a linear interpolation comparing the undiluted CO<sub>2</sub> concentrations in the EUR system to those  
11 from the CHE system was 0.993 (R<sup>2</sup>=0.997). This variability is within the uncertainty of the CO<sub>2</sub>  
12 measurements, estimated to be less than 4% across all test points.  
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21 The exhaust sample for the nvPM line was diluted with particle-free synthetic air (purity  
22 99.999%) using a Dekati DI-1000 ejector diluter. The dilution factors achieved using the DI-1000  
23 ejector diluter are highly dependent on the sample inlet pressure and diluent air pressure  
24 (Giechaskiel, Ntziachristos, & Samaras, 2004). The dilution factor at nominal sample and diluent  
25 inlet pressures was 8-14 (as specified in the SARPs); however, the range extended from 7 to 20  
26 for some of the dilution factor sensitivity tests performed during the campaign. The diluted sample  
27 was transferred to the real-time diagnostic instruments by a 24 - 25 m long, carbon-loaded,  
28 electrically grounded polytetrafluoroethylene (cPTFE) line maintained at a temperature of 60°C ±  
29 15°C followed by a sharp cut cyclone with a 1 µm cut size. The sample lines for each reference  
30 system had slightly different internal diameters, but well within the range 7.59 - 8.15 mm as  
31 specified by AIR6241. The diluted exhaust sample flow rate in the cPTFE line was maintained at  
32 25 slpm ± 2 slpm. Another three-way splitter distributed the particle-laden flow to the nvPM  
33 number instrument, the nvPM mass instruments, and the excess flow line for CO<sub>2</sub> concentration,  
34 and other ancillary measurements that were not a requirement of the standardized system.  
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46 All three reference systems measured nvPM number emissions using an AVL Particle  
47 Counter (APC) Advanced (Giechaskiel et al., 2010), which consists of a primary dilution stage  
48 with a rotating disk diluter, a catalytic stripper with a sulphur trap (volatile particle remover, VPR)  
49 maintained at 350°C, a secondary dilution stage with a porous tube diluter, and an n-butanol-based  
50 condensation particle counter (CPC; TSI 3790E), with a 50% cut-off diameter ≈10 nm and 90%  
51 count efficiency at ≈15 nm (Lobo et al., 2015a). All CPC's were operated in single count mode by  
52 increasing the APC dilution factor when CPC concentrations approached 10,000 particles/cm<sup>3</sup>.  
53 The diluent used for the APC was synthetic air. The nvPM mass emissions were measured using  
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4 both an Artium Laser Induced Incandescence LII 300 (LII) (Snelling et al., 2005) and an AVL  
5 Micro Soot Sensor (MSS) (Schindler et al., 2004) on all three reference systems. Additionally, two  
6 Cambustion DMS500 fast mobility spectrometers (Reavell et al. 2002) were installed on the excess  
7 flow/ancillary line, one each on the EUR and NAM reference systems to measure PM size  
8 distributions. A compact time of flight aerosol mass spectrometer (CToF-AMS) (Drewnick et al.  
9 2005) to obtain chemical composition information, and a Cavity Attenuated Phase Shift (CAPS)  
10 PM extinction monitor (Yu et al., 2011) to measure nvPM mass were installed on the ancillary line  
11 on the NAM reference system. Further aerosol instrumentation was also deployed in the CHE  
12 reference system to measure particle effective density (Durdina et al., 2014) and chemical  
13 composition (Abegglen et al., 2016).  
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### 24 2.4.3 Instrument calibrations

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26 The LII 300 and MSS instruments were calibrated to the NIOSH 5040 protocol (NIOSH,  
27 2003) for EC using thermal optical analysis (TOA) in accordance with AIR6241 in two batches at  
28 NRC-Metrology, Ottawa, Canada one month prior to the start of the campaign. The NRC inverted-  
29 flame burner (Coderre et al., 2011) was used as the source of black carbon (BC) particles. The BC  
30 particles generated by the inverted-flame burner were diluted using filtered air, and the diluted  
31 sample was divided using a splitter and then directed to two 1 $\mu$ m cyclones – one was upstream of  
32 the dual-stage filter collection system, and the other upstream of the nvPM mass instruments to be  
33 calibrated. Equal length (2 m) heated cPTFE 3/8" OD tubing was used from splitters to all  
34 measurement devices. The entire sampling system was heated to 60°C. The MSSs from all three  
35 reference systems and the LII 300 from the NAM reference system were calibrated in the first  
36 batch with 6 or more repeats at target mass concentrations of 0, 50, 100, 500, and 1000  $\mu$ g/m<sup>3</sup>. The  
37 remaining two LII 300s were calibrated in the second batch with 3 or more repeats at target mass  
38 concentrations of 0, 100, 250, and 500  $\mu$ g/m<sup>3</sup>. The correlation of mass concentration measured by  
39 the instruments to NIOSH 5040 EC was >0.995 in all cases.  
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51 The APC for all three reference systems had its annual calibration and maintenance  
52 performed by the manufacturer within nine months of the campaign. Combustion soot from a  
53 miniature combustion aerosol standard (miniCAST) generator was used as the source to establish  
54 penetration through the VPR. CO<sub>2</sub> calibration gas was used to verify dilution factors for the two-  
55 stage dilution of the APC. As part of the calibration procedure, VPR performance was evaluated  
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4 in terms of volatile particle removal efficiency. The CPC linearity and counting efficiency were  
5 also determined and adjusted (counting efficiency at 10 nm > 50%). All three reference systems  
6 had similar particle penetration profiles for the VPR, and volatile particle removal efficiencies of  
7 99.99%. The CHE and NAM reference systems had similar CPC counting efficiencies of 76% and  
8 92% at 10 nm and 15 nm, respectively, while the EUR reference system had a much lower CPC  
9 counting efficiency of 53% at 10nm and slightly higher CPC counting efficiency of 98% at 15 nm.  
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## 18 **2.5 Test Matrix**

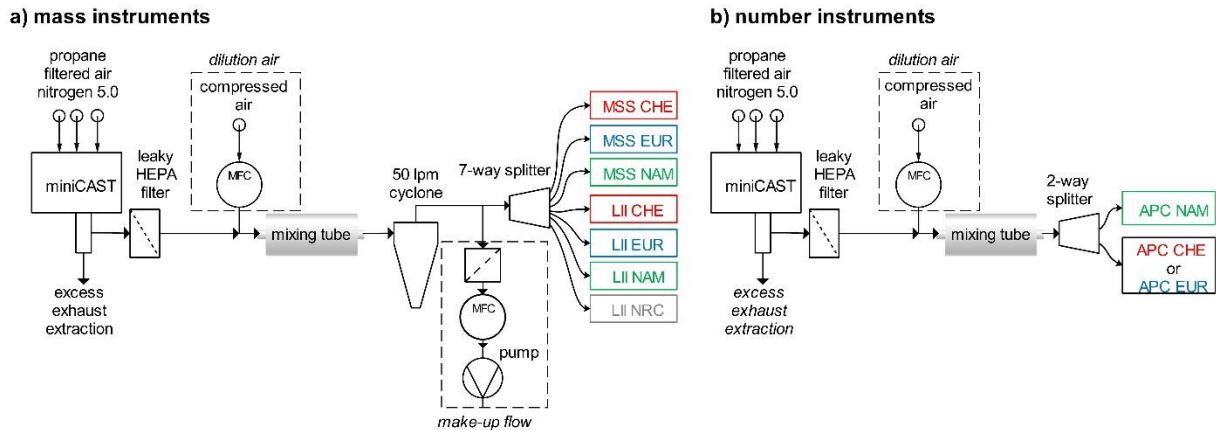
19 The dedicated engine tests with the CFM56-7B26/3 engine started with a warm-up  
20 sequence used for conditioning the probe and the sampling systems. The warm-up sequence  
21 consisted of running the engine at five test points from ground idle to 85% sea-level static thrust  
22 for durations of 5 minutes each. The parameter used for setting the engine thrust was the combustor  
23 inlet temperature, T3. The T3 values were based on a correlation of sea-level static thrust with T3  
24 corrected to ISA conditions (15°C, 1 atm). The full test matrix following the warm-up sequence  
25 consisted of 12 points on a descending power curve, starting at maximum continuous thrust (which  
26 was limited by the ambient conditions during the test) and ending at idle. These test points included  
27 the four thrust settings corresponding to the LTO cycle as well as an additional point at 65%. The  
28 duration of each test point was 10 minutes. A subset of the full test matrix was run during most  
29 dedicated engine tests. Table 2 lists the inter-comparison experiments performed during the  
30 measurement campaign, along with the range of ambient conditions recorded during each test.  
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**Table 2: Overview of experiments performed for Reference Systems inter-comparison**

Date	Ambient Condition Ranges			Engine/Source	Description	Sampling Probe	Reference System		
	T (° C)	P (kPa)	RH (%)				Swiss (CHE)	European (EUR)	North American (NAM)
28 Jul 2013	28.5-34.0	96.0-96.2	24-42	CFM56-7B26/3	AFR check (probe tips 1, 4, 6)	Multi-point	✓	✓	✗
29 Jul 2013	17.3-17.8	96.6-96.8	87-95	CFM56-7B26/3	AFR check (all 24 probe tips)	Multi-point	✓	✓	✗
2 Aug 2013				miniCAST 5201C	Instrument comparison	-	✓	✓	✓
3 Aug 2013	27.4-30.8	96.9-97.0	28-39	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✓	✓
4 Aug 2013	18.0-26.3	97.1-97.5	45-82	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✓	✓
5 Aug 2013	24.2-25.6	96.7-96.8	54-57	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✓	✓
10 Aug 2013	21.1-22.5	97.1-97.2	44-53	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✗	✓
11 Aug 2013	18.4-24.2	96.9-97.1	39-56	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✗	✓
12 Aug 2013	20.4-26.1	96.6-96.8	34-50	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✗	✓
24 Aug 2013	17.0-20.2	96.3-96.5	68-92	CFM56-7B26/3	Reference Systems comparison	Multi-point	✓	✓	✗

## 2.6 miniCAST

A miniCAST soot generator 5201C (Jing Ltd) was used as a surrogate emissions source to compare the nvPM mass and number diagnostic instruments used in the three reference systems prior to the dedicated engine tests. The miniCAST was operated at the following flowrates - propane: 0.06 lpm, N<sub>2</sub> mixing gas: 0 lpm, oxidation air: 1.55 lpm, N<sub>2</sub> quench air: 7 lpm, dilution air: 20 lpm, such that a high elemental carbon (EC) fraction (>80%), determined from TOA, was produced in the exhaust stream (Durdina et al., 2016). The mean size of particles at the miniCAST setting selected was ~130 nm. This test was performed to verify that the operation and performance of the instruments were optimal after transport to the SR Technics engine test facility. The configurations for the nvPM mass and number instrument comparisons using the miniCAST are presented in Figure 2.



**Figure 2: Configuration for the mass (a) and number (b) instrument comparisons using the miniCAST 5201C soot generator**

For the nvPM mass instrument comparison, the miniCAST exhaust was passed through a HEPA filter dilution bridge and then diluted with compressed air to achieve target mass concentrations of 50, 100, 250, 500, 750, and 1000  $\mu\text{g}/\text{m}^3$ . The diluted exhaust sample was then transferred through a mixing tube, a cyclone with a 1  $\mu\text{m}$  cut-point at 50 lpm, and a 7-way splitter to the instruments under test – 6 instruments, (3 MSSs and 3 LII 300s) two from each reference system, and another LII 300 which was used to monitor source concentration levels during the course of the instrument comparisons. Only data from the instruments in the reference systems were used in the analysis. The calibration factors determined during the NRC-Metrology

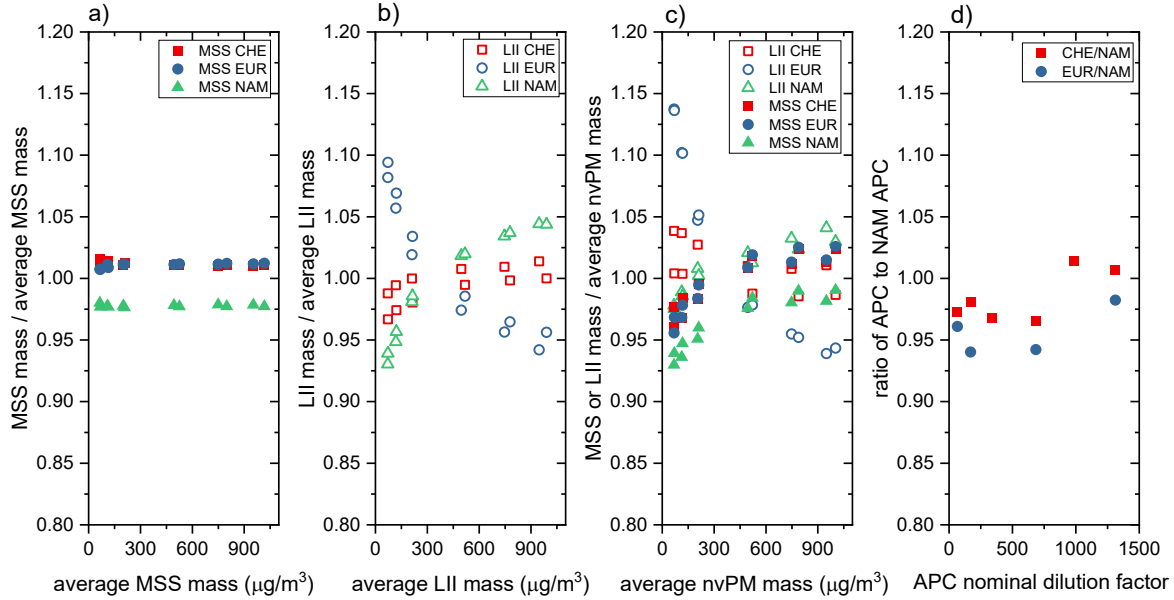


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4 calibrations from the NIOSH 5040 EC protocol for all six nvPM mass instruments were applied  
5 prior to the comparison study. For the nvPM number instrument comparison, the miniCAST  
6 exhaust was diluted with compressed air and split into two legs – one providing a sample to the  
7 APC from the NAM reference system, and the other to the APC from either the CHE or EUR  
8 reference systems.  
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### 14 **3 Results and Discussion**

#### 17 *3.1 Instrument comparisons using the miniCAST*

20 The results of the nvPM mass and number instrument comparisons conducted using the  
21 miniCAST as a source of nvPM are presented in Figure 3. The nvPM mass concentrations as  
22 measured by the MSS (Fig 3a) and LII 300 (Fig 3b) for each reference system were averaged for  
23 60 seconds, and each instrument was then compared against the ensemble average. The data from  
24 the MSS and LII 300 instruments were tightly bound with a relative standard deviation (RSD) of  
25 1.95% and 4.5%, respectively. A relatively higher degree of scatter ( $\pm 10\%$ ) was observed with the  
26 LII 300 data for nvPM mass concentrations below  $100 \mu\text{g}/\text{m}^3$ . Overall, the six nvPM mass  
27 instruments were within an RSD of 4.1% over the range of target nvPM mass concentrations  
28 explored. An RSD of 2.1% was observed between the 3 APCs used for the nvPM number  
29 measurements when the CHE and EUR reference system APCs were compared against the NAM  
30 reference system APC (Fig 3d). The repeatability of the miniCAST as reported by the  
31 manufacturer is  $\pm 5\%$ . Larger differences have been observed from inter-day experiments (Moore  
32 et al. 2014), however, for the instrument comparison test the miniCAST settings were stable with  
33 RSDs in the mean nvPM mass and number concentrations  $< 5\%$ . Since the differences observed  
34 between the instruments were of a similar magnitude, the nvPM mass instruments and the nvPM  
35 number instruments were assessed as being in statistical agreement on the miniCAST source.  
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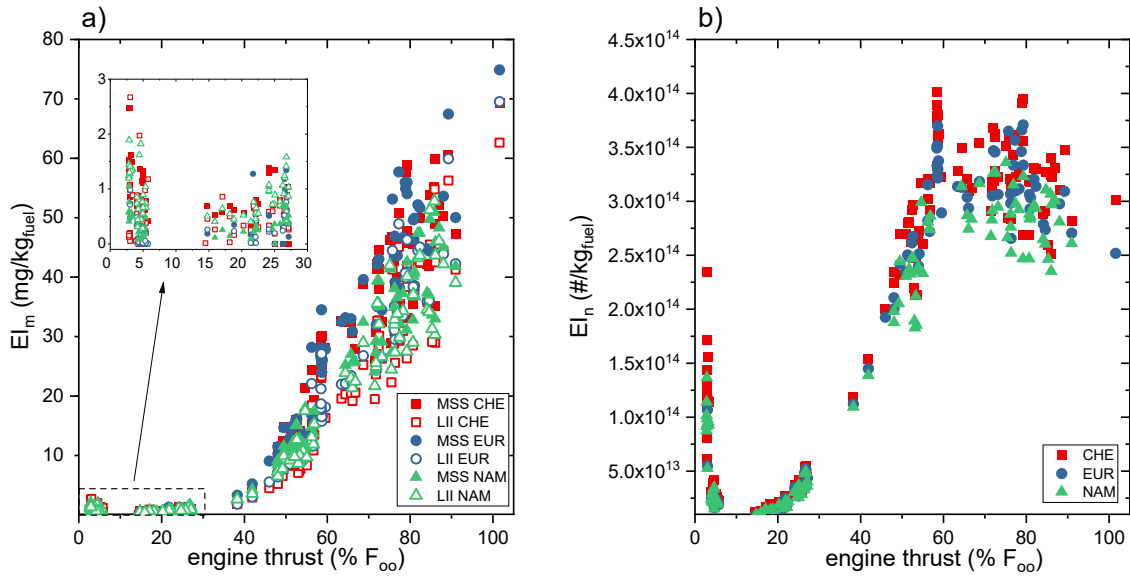


**Figure 3: Comparison of nvPM mass (a, b, c) and number (d) instruments using the miniCAST 5201C soot generator**

### 3.2 *nvPM mass and number emission profiles for the CFM56-7B26/3 engine*

The nvPM mass and number concentrations were converted to nvPM mass-based emission index ( $EI_m$ ) and nvPM number-based emission index ( $EI_n$ ), respectively, using the measured nvPM and gaseous emissions concentration and following the procedures specified in AIR6241 (SAE, 2013). The EIs are reported at a standard temperature of 273.15 K and standard pressure of 101.325 kPa. The nvPM mass and number emission indices were not corrected for either the thermophoretic loss in the sample extraction system or for size-dependent diffusional and inertial losses that occurred in the sampling and measurement systems, and CPC efficiencies for nvPM number. The nvPM mass and number emission indices for the CFM56-7B26/3 engine as a function of percent rated thrust are presented in Figure 4. For the nvPM mass emission index profile, the emissions for this engine were generally higher at idle conditions (3-7% rated thrust), decreased to a minimum at low engine thrust conditions (15-30% rated thrust), and increased linearly to maximum rated thrust. The nvPM number emission index also exhibited behavior similar to that of nvPM mass emissions at idle and low engine thrust conditions. However, the emissions increased up to a maximum at ~60% rated thrust and then slightly decreased up to the maximum rated thrust. The nvPM mass and number emissions profiles shown here are consistent with

previously reported profiles for a CFM56-7B24/3 engine (which is the same engine model but rated at a lower take-off thrust) (Lobo et al., 2015a).



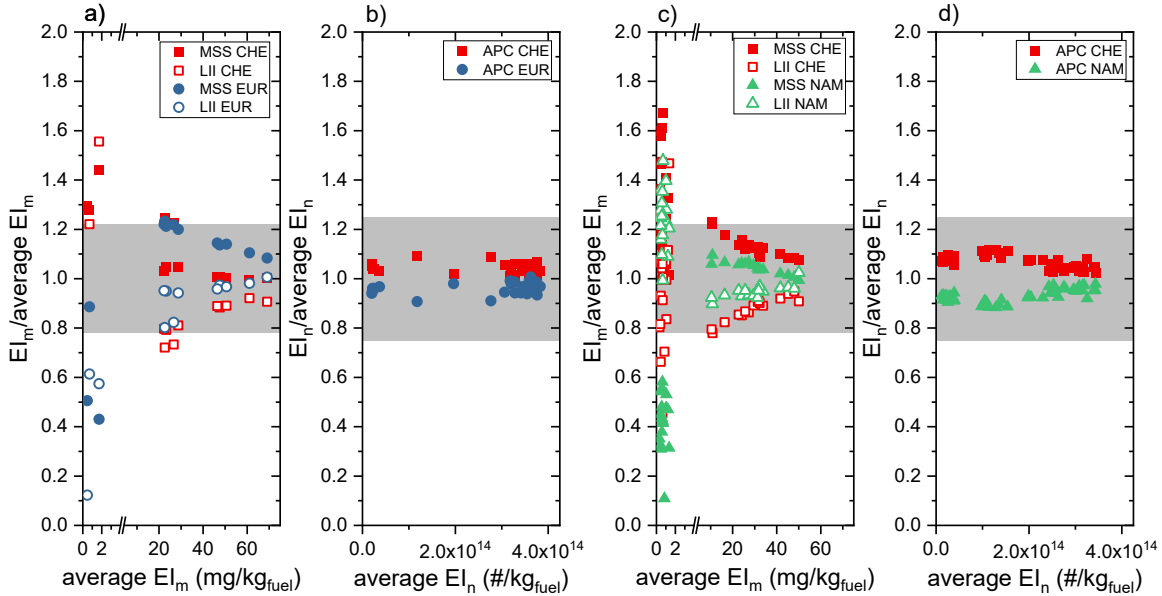
**Figure 4: nvPM mass (a) and number (b) emission index profiles for the CFM56-7B26/3 engine**

### 3.3 2-way reference system comparisons (CHE-EUR; CHE-NAM)

Comparison of nvPM mass and number emissions for pairs of reference systems, i.e., CHE-EUR and CHE-NAM was performed during the test campaign and is presented in Figure 5. The multi-point sampling probe was used for the CHE-EUR comparison, while the single-point probe was used for the CHE-NAM comparison. However, since the extracted exhaust samples were representative of the engine emissions, the use of a particular sampling probe did not influence the comparisons between the reference systems. The 2-way comparisons with reference systems in parallel were performed to assess system to system differences without the additional complexity of including a third system (also performed and described in the next section). The nvPM mass and number emission index for each instrument in the reference system pairs were averaged and compared against the ensemble average for a specific pair-wise comparison. The average nvPM mass and number emission indices for the CHE-EUR comparison were slightly higher than the CHE-NAM comparison because of the higher engine thrust conditions achieved during the respective tests.

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4 The theoretical total uncertainties in the nvPM EIs presented are estimated to be ~22% for  
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6 EI mass and ~25% for EI number. These estimates are based on typical uncertainty values for the  
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8 nvPM mass and number instruments (including calibration uncertainty), CO<sub>2</sub> measurement, and  
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10 the determination of the dilution factor in the nvPM number instrument (SAE, 2013). These  
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12 estimates do not account for either the particle losses in the sampling and measurement systems or  
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14 the increased uncertainty for nvPM mass measurements near the limit of detection (LOD).  
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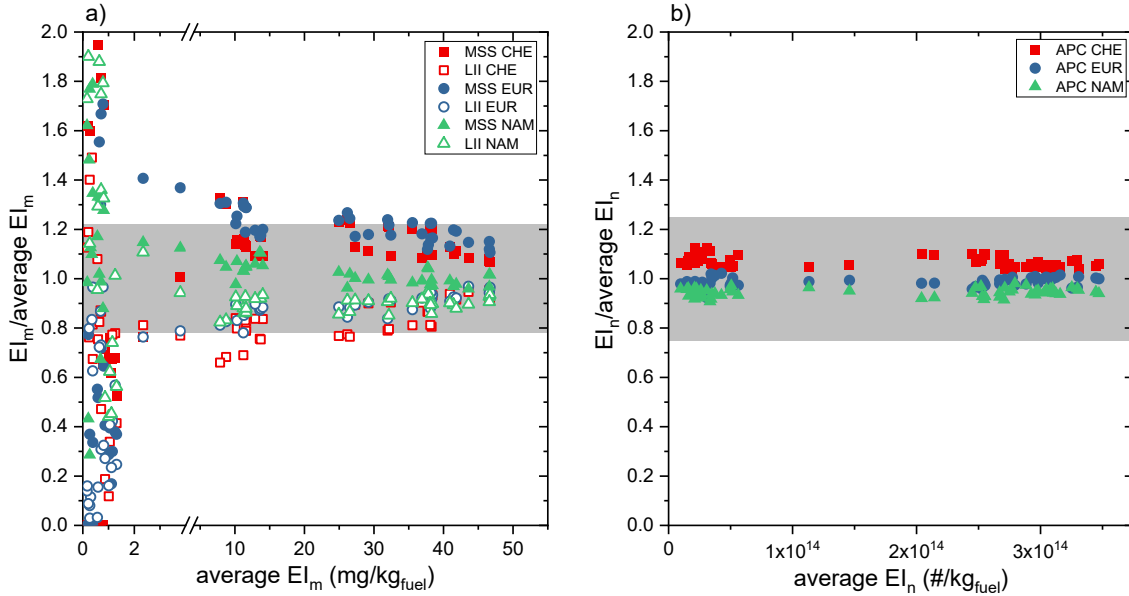
19 All instruments in the three reference systems were generally ±22% of the average nvPM  
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21 mass emission index and ± 25% of the average nvPM number emission index (grey shaded area  
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23 in Fig. 5), except for the nvPM mass emission index which exhibited a >22% variation at low  
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25 nvPM mass concentrations levels (corresponding to EIs of < 30 mg/kg), and significantly higher  
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27 differences as the instruments approached the LOD (3 µg/m<sup>3</sup>, corresponding to EIs of <2 mg/kg).  
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29 The high variability in mass at low concentrations was exhibited for both types of nvPM mass  
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31 instruments (MSS and LII 300). This trend for high variability in nvPM mass at low concentrations  
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33 is consistent with results for emissions measurements of other engine types (Lobo et al., 2015a;  
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35 Lobo et al., 2016). For nvPM number, the CHE reference system registered consistently higher  
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37 values compared to the EUR and NAM reference systems. The overall magnitude of variation for  
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39 each instrument from the average was consistent for the two pairs of reference system  
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41 comparisons, i.e ± 15%. It should be noted that since only a single type of nvPM number  
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43 instrument was compared, there is no information on the uncertainty associated with using  
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45 different types of nvPM number instruments that meet the specifications.  
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**Figure 5: Comparison between CHE and EUR reference systems (a, b) and CHE and NAM reference systems (c, d) for nvPM mass and number emission indices**

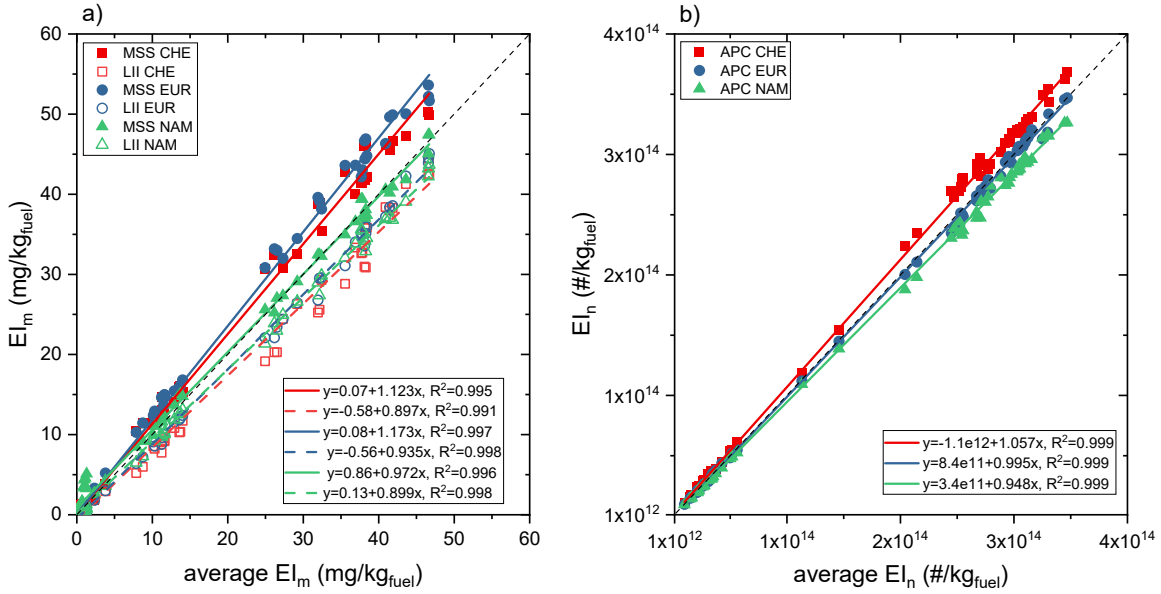
### 3.4 3-way reference system comparisons (CHE-EUR-NAM)

All three reference systems were compared simultaneously using the CFM56-7B26/3 engine. The comparisons between the reference systems in terms of the ratio of the nvPM mass and number emission index to the average emission index as a function of the average emission index are presented in Figure 6. Similar trends in nvPM mass and number EIs during the 2-way comparisons are also observed for the 3-way comparisons. The average nvPM mass and number EIs for the 3-way reference system comparisons are lower than those for the 2-way comparisons because the nvPM emissions produced by the engine for these tests were lower. Variability in nvPM mass emission index was higher than 20% for EIs up to 40 mg/kg fuel (corresponding to a mass concentration of  $\sim 95 \mu\text{g}/\text{m}^3$  at the instrument), while the variability in nvPM number emission index was  $\pm 15\%$  for all test conditions.



**Figure 6: Comparison between CHE, EUR, and NAM reference systems for nvPM mass (a) and number emission index (b).**

It is also informative to view the nvPM mass and number emission indices for the instruments used in the three reference systems using parity plots as shown in Figure 7. The EI data reported for each instrument are plotted against the average EI. While these plots are not suitable to illustrate differences at low concentration levels, they provide an overall magnitude of variability between the instruments. As can be seen in Figure 7, the nvPM  $EI_m$  and  $EI_n$  for each instrument were well correlated with the average. The nvPM  $EI_n$  for all three systems was within  $\pm 6\%$  of the average. For the nvPM  $EI_m$ , the magnitude of the differences was  $\sim 10\%$  for the LII 300 and  $\sim 15\%$  for the MSS. Ideally, quartz filters would have been collected in parallel with the real-time instruments to determine EC content using TOA, and then used as the reference to compare the nvPM mass instruments. The filter collection for TOA was not performed during the campaign due to limitations on sampling time at each test condition preventing adequate sample to be collected for analysis.



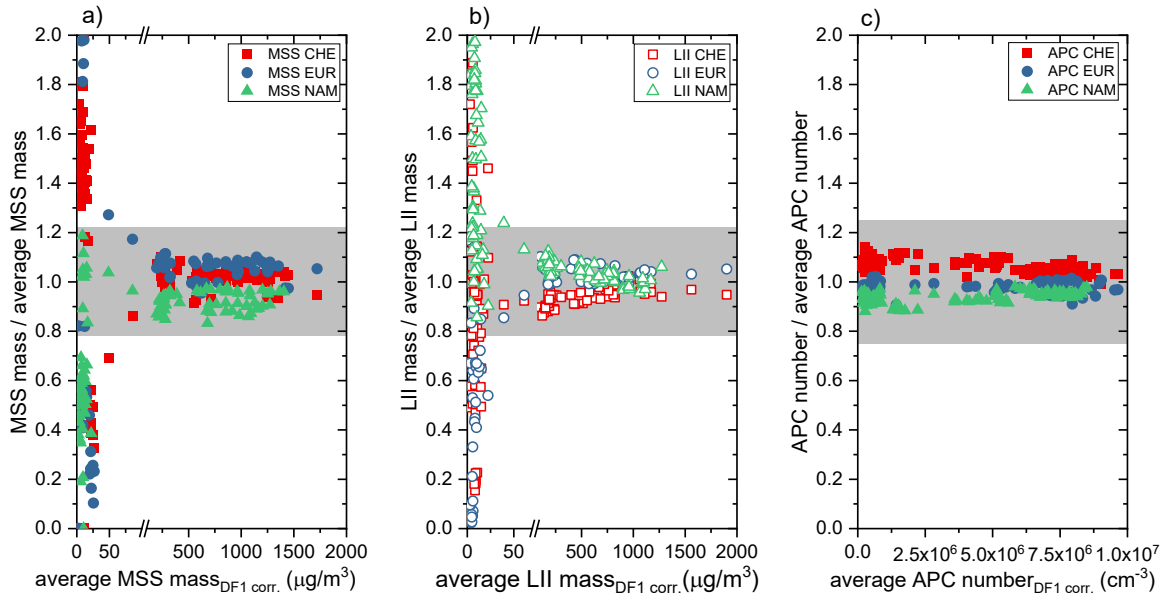
**Figure 7: Parity plot comparisons of CHE, EUR, and NAM reference systems for nvPM mass (a) and number emission index (b).**

### 3.5 Reference system comparisons for nvPM mass and number concentration

The comparison between the reference systems in terms of nvPM mass and number concentrations for each type of measurement instrument is presented in Figure 8 to assess performance on a concentration basis (the primary output of the instrument). The concentration data at the measurement location (Figure 8) have been corrected for dilution since each system had slightly different dilution factors. All data recorded during the campaign when at least 2 systems were operating in parallel are included in this analysis. The comparison between the three reference systems is presented as a function of the ensemble averages for the different types of measurement instruments, i.e. MSS and LII 300 for nvPM mass, and APC for nvPM number.

As was previously reported, the largest differences between the three reference systems for nvPM mass were observed for dilution corrected nvPM mass concentrations  $< 50 \mu\text{g}/\text{m}^3$  at the measurement location ( $\sim 5 \mu\text{g}/\text{m}^3$  at the instrument). Beyond this threshold, the particular nvPM mass instrument type, i.e. MSS or LII 300, in the reference systems were within 20% of the instrument-specific average mass concentrations. For a given nvPM mass instrument type, the variability in the measured nvPM mass emissions is constrained in a narrow range, which is not the case when both nvPM mass instrument types are included in the analysis (see Figures 5 and

6). Unlike the instrument comparisons with the miniCAST (Figures 3 a and b), both types of nvPM mass instruments each demonstrate similar variability, exceeding 20% only below 50  $\mu\text{g}/\text{m}^3$  (Figures 8 a and b) on engine exhaust. For nvPM number, all three reference systems were well within 20% of the dilution corrected average concentration over the entire range of values recorded.



**Figure 8: Comparison between CHE, EUR, and NAM reference systems for nvPM mass (a,b) and number concentrations (c). The concentrations are reported at the measurement location and corrected for dilution.**

### 3.6 Variability

The variability in nvPM mass and nvPM number emissions was computed by calculating the RSD of the ratio of the respective EI to the average EI (data from Figures 5 and 6). This method for determining variability was adopted to focus on the relative response of the instruments (as a function of concentration), and to decouple the thrust dependency of the EIs and variability in ambient temperature. The measurement campaign was conducted over the course of a month, and the wide range of ambient conditions affected the nvPM emissions produced by the engine.



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4 Previous studies have also reported on the impact of ambient conditions such as temperature on  
5 nvPM emissions variability (Lobo et al., 2015a).  
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8 The three reference systems were inter-compared to establish repeatability and  
9 intermediate precision of the sampling and measurement systems. Repeatability is defined as the  
10 variability of many measurements where the same equipment and operator are used to make  
11 repeated measurements over a short time period, while intermediate precision refers to the  
12 variability of measurements when only some of the four precision conditions (time, calibration,  
13 equipment, operator) are different (JCGM, 2012). The variability in nvPM mass emissions for  
14 repeatability (intra-system) and intermediate precision (inter-system) comparisons are presented  
15 in Table 3 as a function of average nvPM concentrations, with lowest concentrations of nvPM  
16 mass and number grouped in the case of the CFM56-7B26/3 engine at low engine thrust ranges(3-  
17 30%), with increasing concentrations averaged at medium engine thrust (38-60%), and high engine  
18 thrust (63-101%). A similar analysis for nvPM number emissions for intermediate precision  
19 comparisons is presented in Table 4. Intra-system (repeatability) variability for nvPM number  
20 emissions is not considered since each reference system used the same instrument type (APC) for  
21 the measurement.  
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25 The variability for nvPM mass was highest in the lowest mass concentrations (low engine  
26 thrust) range, where the average instrument concentration was below the LOD for both types of  
27 nvPM mass instruments at  $0.7 \mu\text{g}/\text{m}^3$ . The resolution of the MSS instruments used during the  
28 campaign was  $1 \mu\text{g}/\text{m}^3$ , whereas the LII 300 had a resolution of  $0.01 \mu\text{g}/\text{m}^3$ . The higher variability  
29 of the MSS compared to LII 300 at the LOD is likely introduced through the resolution of the  
30 instrument. For medium and high concentrations (successively higher engine thrust ranges), the  
31 variability within a system and between the reference systems was  $<13\%$ . For nvPM number, the  
32 variability was  $<3\%$  across the engine thrust ranges. The sensitivity of nvPM number to the limit  
33 of detection was not a factor since the measured concentration was significantly above the LOD.  
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**Table 3: Variability in nvPM mass emissions**

Comparison	Average instrument mass concentration	0.7 µg/m <sup>3</sup>						28 µg/m <sup>3</sup>						87.8 µg/m <sup>3</sup>					
	Thrust range	Low (3-30%)						Medium (38-60%)						High (63-101%)					
	Reference System	CHE MSS	EUR MSS	NAM MSS	CHE LII	EUR LII	NAM LII	CHE MSS	EUR MSS	NAM MSS	CHE LII	EUR LII	NAM LII	CHE MSS	EUR MSS	NAM MSS	CHE LII	EUR LII	NAM LII
Repeatability (Intra-system)	CHE	62%			48%			6.9%			10%			5.5%			7.1 %		
	EUR		65%			71%			3.8%			5.7%			3.7%			4.8%	
	NAM			60%			33%			3.1%			3.7%			2.4%			2.6%
Intermediate precision (Inter-system)	CHE-EUR and CHE-NAM	16%	40%	36%	32%	62%	30%	8.7%	1.2%	1.8%	4.6%	8%	2%	5.8%	4.1%	2.2%	5.1%	6.9%	2.7%
	CHE-EUR-NAM	126%	109%	68%	71%	75%	56%	12.8%	5.2%	5.5%	8%	5.3%	7.9%	5.2%	3.2%	5.6%	8%	3.7%	2.4%

**Table 4: Variability in nvPM number emissions**

Comparison	Average instrument number concentration (corrected for 2-stage dilution in the APC)	4.9×10 <sup>4</sup> /cm <sup>3</sup>			5.4×10 <sup>5</sup> /cm <sup>3</sup>			7.6×10 <sup>5</sup> /cm <sup>3</sup>		
	Thrust range	Low (3-30%)			Medium (38-60%)			High (63-101%)		
	Reference System	CHE	EUR	NAM	CHE	EUR	NAM	CHE	EUR	NAM
Intermediate precision (Inter-system)	CHE-EUR and CHE-NAM	2.1%	2.3%	1.8%	2.2%	2.2%	1.7%	1.8%	3.1%	1.0%
	CHE-EUR-NAM	2.4%	2.7%	1.7%	2.0%	1.1%	2.4%	1.0%	1.3%	0.9%

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4 The other significant contributor to the variability in nvPM EI<sub>m</sub> and nvPM EI<sub>n</sub> is the CO<sub>2</sub>  
5 concentration. The diluted CO<sub>2</sub> measurements were used to calculate the EIs from mass and  
6 number concentrations measured by each reference system. It was not possible to evaluate the  
7 variability of the diluted CO<sub>2</sub> measurements. Each reference system had slightly different  
8 dimensions for the ejector-diluter vent, which resulted in subtle differences in overall dilution  
9 factors. A comparison of the CO<sub>2</sub> analyzers measuring the same exhaust sample on the diluted  
10 nvPM line during the engine tests was not performed. However, the variability in undiluted CO<sub>2</sub>  
11 concentrations measured by the CHE and the EUR reference systems was evaluated. The slope of  
12 a linear interpolation comparing the undiluted CO<sub>2</sub> concentrations in the EUR system to those  
13 from the CHE system was 0.993 (R<sup>2</sup>=0.997).  
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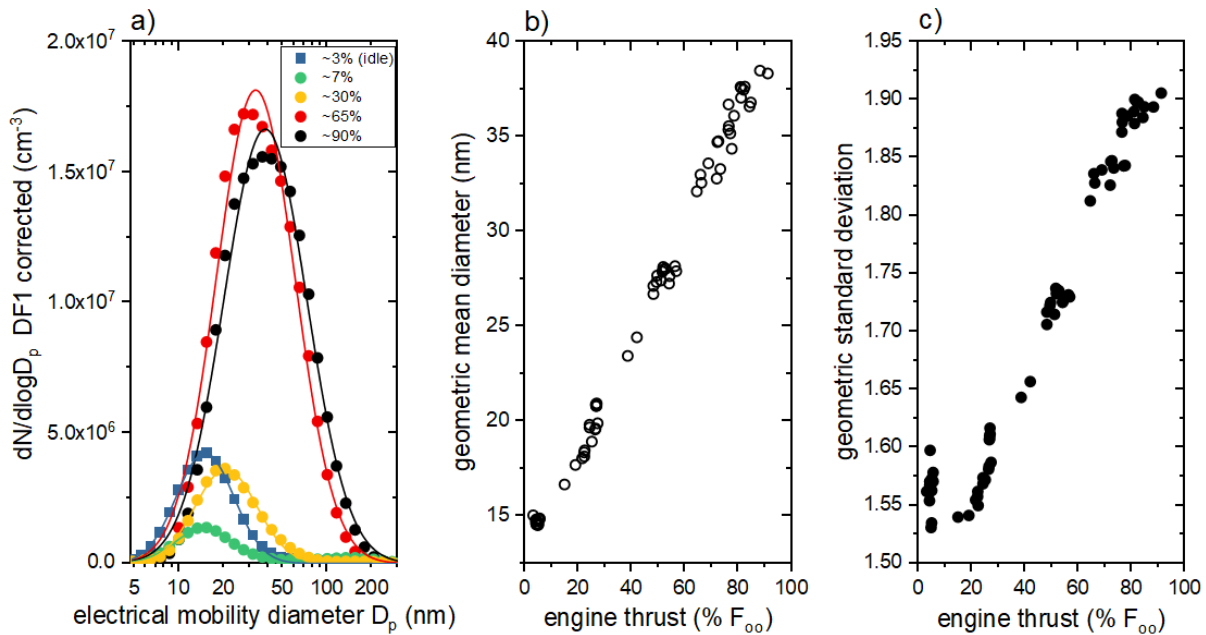
22 All three reference systems were built to and compliant with the specifications for the  
23 standardized system detailed in AIR6241, and in this case used nominally identical nvPM  
24 instruments. Hence, the differences in particle losses in these three sampling and measurement  
25 systems are expected to be negligible compared to the variability in other factors described  
26 previously.  
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32 Building on the knowledge gained from this campaign, several changes to the instrument  
33 performance and calibration protocols were implemented. The resolution of the MSS instruments  
34 was updated to 0.01 µg/m<sup>3</sup>. The procedure to demonstrate conformity of the nvPM mass  
35 instruments to performance specifications was updated to include an additional applicability  
36 criterion for validation of the calibration to EC on aircraft turbine engine exhaust. The limit of  
37 detection of the nvPM mass instruments was also lowered from 3 µg/m<sup>3</sup> to 1 µg/m<sup>3</sup> (ICAO, 2017).  
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### 44 **3.7 Size distributions**

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46 The standardized protocol for aircraft engine nvPM mass and number emissions does not  
47 specify a measurement of particle size distribution. However, a size distribution measurement is  
48 being considered for future standardized methodologies for particle loss correction. Since particle  
49 loss mechanisms such as diffusion and inertial losses are size-dependent, measurement of size  
50 distributions along with nvPM number and mass concentration provides information to estimate  
51 particle loss factors. These loss factors can then be used to calculate nvPM emissions at the engine  
52 exit plane. Engine exit plane emissions would be more relevant for aircraft engine nvPM emissions  
53 inventory and impact assessments. The size distributions for the CFM56-7B26/3 engine along with  
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characteristic parameters – geometric mean diameter (GMD) and geometric standard deviation (GSD) are presented in Figure 9 as a function of engine thrust setting. These size distributions were obtained with the DMS500 installed on the ancillary line of the EUR reference system and corrected for primary dilution (DF1) in the ejector diluter. The size distributions could be approximated to lognormal distributions ( $R^2 > 0.97$ ) with GMD ranging from 15nm at idle to 38nm at 90% rated thrust, and GSD varying between 1.53 and 1.92. The magnitude and general increasing trend of GMD and GSD with engine thrust setting are consistent with previously reported values for this engine type (Lobo et al., 2011; Lobo et al., 2015a; Durdina et al., 2017; Elser et al., 2019).



**Figure 9: Particle size distributions (a) and characteristic parameters – GMD (b) and GSD (c) for the CFM56-7B26/3 engine**

#### 4 Conclusions

Three reference systems for aircraft engine nvPM emissions measurement – the Swiss (CHE) system, the European (EUR) system, and the North American (NAM) system – were developed in compliance with the specifications for the standardized sampling and measurement methodology. The first and only inter-comparison to date of these three reference systems was performed at the SR Technics engine test facility in Zürich, Switzerland using a commercial CFM56-7B26/3 aircraft engine as the emissions source to establish repeatability and intermediate

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4 precision of the sampling and measurement systems. All three reference systems measured nvPM  
5 number concentration using an APC, and nvPM mass concentration was measured using both an  
6 LII 300 and an MSS. The nvPM mass and number concentrations were converted to their  
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8 respective emission indices for comparison. The specifications for the standardized sampling and  
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10 measurement system implemented in the three reference systems were robust, as demonstrated by  
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12 the variability observed between the systems. During the dedicated engine tests with the CFM56-  
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14 7B26/3 engine, all instruments in the three reference systems were generally within 30% of the  
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16 average nvPM mass emission index (determined with different nvPM mass instrument types and  
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18 manufacturers) and 15% of the average nvPM number emission index (determined with the same  
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20 nvPM number instrument type and manufacturer) (see Fig. 6). The only exception was for the  
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22 mass instruments, which exhibited a higher variation as the concentration levels approached the  
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24 LOD of 3  $\mu\text{g}/\text{m}^3$ . A comparison between the three reference systems as a function of the  
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26 measurement instrument type revealed that similar measurement methodologies had a better  
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28 agreement and lower variability. As more fuel efficient aircraft engines with low emission  
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30 combustors continue to be developed, instruments for measuring nvPM mass should have the  
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32 capability of higher resolution and sensitivity for low concentration levels. Future studies should  
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34 consider the variability associated with other instruments that meet the performance specifications  
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36 in AIR6241 but were not evaluated in this study.

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38 It should be noted that the emission index values reported for nvPM mass and number have  
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40 not been corrected for size-dependent particles losses in the sampling and measurement systems,  
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42 and hence do not represent the actual emissions at the engine exit plane. Including a traceable size  
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44 measurement in the standardized measurement system would enable a more accurate estimation  
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46 of engine exit plane nvPM emissions to improve airport emissions inventory development and  
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48 environmental impact assessment of aircraft engine nvPM emissions. Size distribution  
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50 measurements, not currently specified in the standard method, were found to be approximated to  
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52 lognormal distributions with GMD ranging 15nm - 38nm, and GSD varying 1.53 - 1.92.

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54 The wide range of ambient conditions encountered during the campaign affected the nvPM  
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56 emissions produced by the engine. A correction for changes in ambient conditions will need to be  
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58 developed to decouple the variability in the ambient temperature from the measured nvPM mass  
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60 and number emissions. Although the CFM56-7B26/3 engine used in this study is the most widely  
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62 used engine in commercial aviation, other engine types could have different emissions profiles. It  
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4 is essential that the reference systems be compared using an aircraft engine source with a different  
5 emissions profile to validate the repeatability and intermediate precision of the sampling and  
6 measurement systems established in this study. Also, long term comparison of the reference  
7 systems should be undertaken since these systems will continue to be used to varying extents over  
8 time.  
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13 As a direct consequence of the results from this campaign, several changes to the instrument  
14 performance were implemented such as updating the resolution of the MSS instruments to 0.01  
15  $\mu\text{g}/\text{m}^3$ , and lowering the limit of detection of the nvPM mass instruments from 3  $\mu\text{g}/\text{m}^3$  to 1  $\mu\text{g}/\text{m}^3$ .  
16 The procedure to demonstrate the conformity of the nvPM mass instruments to performance  
17 specifications was updated to include an additional applicability criterion for validating the nvPM  
18 mass instrument calibration on aircraft turbine engine exhaust.  
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24 The results from this study are a benchmark for the variability in standardized sampling  
25 and measurement systems for measuring aircraft engine emissions. The three reference systems  
26 evaluated were subsequently used for comparisons with aircraft engine manufacturer sampling and  
27 measurement systems. The aircraft engine manufacturers contributed nvPM emissions datasets for  
28 24 aircraft engine types that were representative of the current commercial fleet for inclusion in a  
29 database (Agarwal et al., 2019). With the database and knowledge of the uncertainty as  
30 characterized by the intermediate precision, the new ICAO nvPM mass and number emissions  
31 regulatory standard for in production and new engines (CAEP/11) was developed. This new ICAO  
32 regulatory standard will be used to certify all aviation engines with rated thrust > 26.7 kN for  
33 nvPM mass and number emissions performance.  
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4 **Comparison of Standardized Sampling and Measurement**  
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7 **Reference Systems for Aircraft Engine**  
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9 **Non-volatile Particulate Matter Emissions**  
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## 10 **Abstract**

11 The International Civil Aviation Organization has established new regulatory standards for  
12 emissions certification of non-volatile particulate matter (nvPM) from aircraft turbine engines. The  
13 adoption of the nvPM emissions regulatory standards required development of a standardized  
14 sampling and measurement methodology, and rigorous testing. Three reference systems for aircraft  
15 engine nvPM emissions measurement, compliant with the specifications for the standardized  
16 methodology, were independently developed. This paper reports the results of the first inter-  
17 comparison of these three reference systems using a CFM56-7B26/3 aircraft engine to establish  
18 repeatability and intermediate precision of the sampling and measurement systems as part of the  
19 multi-agency international collaborative projects: Aviation-Particle Regulatory Instrumentation  
20 Demonstration Experiment (A-PRIDE) 5/ Studying, sAmpling and Measuring of aircraft  
21 ParticuLate Emissions (SAMPLE) III - SC03. The instruments used in the three reference systems  
22 recorded nvPM mass and number concentration, which were converted to their respective emission  
23 indices for comparison. The reference systems generally agreed to within 15% of the average  
24 nvPM number emission index and 30% of the average nvPM mass emission index. The only  
25 exception was for the nvPM mass instruments, which exhibited a higher variation as the  
26 concentration levels approached the limit of detection. The additional measured particle size  
27 distributions could be approximated to lognormal distributions with the geometric mean diameter  
28 ranging from 15 nm to 38 nm, and the geometric standard deviation varying between 1.53 and  
29 1.92. The results from this study are a benchmark for the variability in standardized sampling and  
30 measurement systems for measuring aircraft engine nvPM emissions.  
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## 50 **Highlights**

- 51 • Comparison of three reference systems for measuring aircraft engine nvPM emissions
  - 52 • The nvPM number emission index was generally within 15% of the average value
  - 53 • The nvPM mass emission index was generally within 30% of the average value
  - 54 • Mass-based emissions exhibited high variability towards the instrument limit of detection
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4 **Keywords**

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6 Aircraft engines, non-volatile particulate matter, aviation emissions, black carbon, particle  
7 number, size distributions  
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11 **1 Introduction**

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13 Aircraft engine gaseous and particulate matter (PM) emissions are a unique source of  
14 pollution compared to other sources in the urban environment. The public awareness about  
15 aviation emissions has grown since the rapid increase in demand for commercial air travel in the  
16 1960s, which led to the introduction of emission standards by the International Civil Aviation  
17 Organization (ICAO). Prior to 2010, the ICAO Committee on Aviation Environmental Protection  
18 (CAEP) developed emission standards and recommended practices (SARPs) limited to the  
19 emissions of gaseous pollutants such as oxides of nitrogen (NO<sub>x</sub>), carbon monoxide (CO), and  
20 unburnt hydrocarbons (UHC), and emissions of smoke (reported in terms of smoke number, SN).  
21 The SARPs are applicable to turbojet and turbofan engines with maximum sea-level static rated  
22 thrust >26.7 kN for gaseous emissions and to all engine sizes for smoke emissions. The SARPs,  
23 intended to mitigate the impact of aircraft engine emissions on local air quality, have been  
24 established for the type certification of aircraft engines and are documented with approved test and  
25 measurement procedures in ICAO Annex 16 to the Convention on International Civil Aviation,  
26 Volume II (ICAO, 2017). The type certification process involves operating one or multiple  
27 representative engines of a specific model on a test stand at combustor inlet temperatures  
28 corresponding to the four thrust settings of the standardized Landing and Take-off (LTO) cycle –  
29 7% (taxi), 30% (approach), 85% (climb), and 100% (take-off) – corrected to International Standard  
30 Atmosphere (ISA) conditions (ICAO, 1993). The engine manufacturers submit emissions data,  
31 acquired during these engine type certification tests to the certifying authority for approval and  
32 subsequently for inclusion in the ICAO Aircraft Engine Emissions Databank (EEDB) maintained  
33 by the European Union Aviation Safety Agency (EASA) (EASA, 2019a).  
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52 Since aircraft engine emissions regulatory standards were first adopted in 1981,  
53 ICAO/CAEP has developed increasingly stringent standards for NO<sub>x</sub> emissions. This, along with  
54 the introduction of newer, more fuel-efficient engine technologies, has resulted in lower aircraft  
55 engine emissions over time (Wey & Lee, 2018). However, with the growth of commercial and  
56 cargo air traffic at a rate of ~4.5% per year (Airbus, 2018; Boeing, 2018; IATA, 2019) the absolute  
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4 emissions from the aviation sector have increased, including PM emissions which are forecast to  
5 increase over the next twenty years if current engine technology continues to be used (EASA,  
6 2019b). This in turn has led to several studies to measure, model, and ultimately develop solutions  
7 to mitigate aviation-related emissions (Masiol and Harrison, 2014).  
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11 The smoke standard was introduced at a time when the visibility of exhaust plumes from  
12 aircraft engines was of significant concern for airport safety. Smoke number is determined by  
13 measuring the reduction in reflectance of a Whatman #4 filter used to sample a prescribed mass of  
14 exhaust per unit area of the filter, i.e. 16.2 kg of exhaust gas/m<sup>2</sup>. The precision of the SN  
15 measurement method is reported as ±3 SN (SAE, 2011), which in some cases is higher than the  
16 recorded SN for modern turbofan engines. Also, the SN does not provide any information about  
17 the number, size, and composition of the PM emissions that are required for environmental impact  
18 assessments. In the absence of fleet-wide aircraft engine PM emissions data, the First Order  
19 Approximation (FOA) methodology was developed based on correlations between SN reported in  
20 the ICAO EEDB and available data for non-volatile particulate matter (nvPM) mass emissions  
21 (Wayson, Fleming, & Iovinelli, 2009). FOA and subsequent updates (current version FOA 4.0)  
22 have been used to estimate PM emissions from certified commercial aircraft engines during the  
23 LTO cycle at airports (Rissman et al., 2013; Winther et al., 2015; Woody et al., 2016).  
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27 The eighth meeting of CAEP (ICAO, 2010) recognized the need for PM emissions  
28 regulation in addition to the existing set of regulated pollutants. The first new regulatory standard  
29 for aircraft engine nvPM emissions was adopted by CAEP in 2016 (CAEP/10). It included  
30 reporting requirements and a nvPM mass emissions standard for in-production aircraft engines  
31 with rated thrust >26.7 kN on or after 1 January 2020. The CAEP/10 standard was set at a  
32 regulatory level that matched the existing smoke number visibility standard. The regulatory level  
33 for the CAEP/10 maximum nvPM mass concentration was developed based on a statistical  
34 relationship between nvPM mass concentration and smoke number (Agarwal et al., 2019). This  
35 ensured that any aircraft engine that would have met the certification requirements for smoke  
36 number would also meet the nvPM mass emissions level (all-pass standard). From 1 January 2023,  
37 regulatory limits for mass and number-based nvPM emissions will become applicable for in-  
38 production and new commercial aircraft engine types with a maximum rated thrust > 26.7 kN  
39 (CAEP/11 standard) (ICAO, 2019). Along with this new standard and since the CAEP/10 standard  
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4 includes control of exhaust plumes visibility, CAEP agreed that the smoke number standard will  
5 no longer be applicable for these engines from 1 January 2023.  
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8 Before regulatory standards for nvPM mass and number emissions could be defined, a  
9 standardized sampling and measurement methodology was required for aircraft engine emissions  
10 certification tests. The Society of Automotive Engineers (SAE) Aircraft Engine Gas and  
11 Particulate Emissions Measurement Committee (E-31) was tasked with recommending a  
12 standardized protocol for aircraft engine nvPM mass- and number-based emissions. The  
13 development of this standardized protocol for aircraft engine nvPM emissions was the culmination  
14 of several years of effort to investigate sampling methods, evaluate measurement technologies,  
15 and assess engine type and fuel composition differences on nvPM emissions during the  
16 PARTEMIS (Petzold et al., 2003), APEX (Lobo et al., 2007; Timko et al., 2010; Kinsey et al.,  
17 2010), AAFEX (Kinsey et al., 2012), SAMPLE (Petzold et al., 2011; Crayford et al., 2012;  
18 Crayford et al., 2013; Boies et al., 2015), A-PRIDE (Durdina et al., 2014; Lobo et al., 2015a; Brem  
19 et al., 2015), and MERMOSE (Delhaye et al., 2017) projects, in addition to other studies (Lobo et  
20 al., 2015b). The results informed the development and publication of Aerospace Information  
21 Report (AIR) 6241 (SAE, 2013) and Aerospace Recommended Practice (ARP) 6320 (SAE, 2018).  
22 The standardized protocol is limited to only nvPM emissions, defined as particles that exist at the  
23 aircraft engine exhaust nozzle exit plane that do not volatilize at temperatures greater than 350°C  
24 (ICAO, 2017). Controlling aircraft engine nvPM emissions at the source will lead to lower  
25 emissions on local, regional, and global scales. Total PM emissions downstream of aircraft engines  
26 are not currently considered since they have been shown to vary in space and time as the exhaust  
27 plume cools and expands (Lobo et al., 2012; Timko et al., 2013; Beyersdorf et al., 2014), making  
28 it complicated to develop a standardized sampling and measurement protocol that reports  
29 repeatable concentrations, and to enforce through a certification requirement. The approach to  
30 limit the sampling and measurement system to nvPM was similar to that adopted by the Particle  
31 Measurement Programme (PMP) for the regulation of the number concentration of solid (non-  
32 volatile) particles with a diameter >23 nm emitted from automotive engines (Giechaskiel et al.,  
33 2012).  
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55 Analogous to the reference “golden” particle number measurement system used in PMP  
56 (Martini, Giechaskiel, & Dilara, 2009), three reference systems for aircraft engine nvPM emissions  
57 measurement were independently developed - the Swiss (CHE) fixed (later mobile) system by  
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4 Empa, the European (EUR) mobile system by Cardiff University, and the North American (NAM)  
5 mobile system by Missouri University of Science and Technology. All three reference systems  
6 were compliant with the specifications for the standardized system detailed in AIR6241, but not  
7 identical since AIR6241 permits tolerances for different components and specification ranges. It  
8 was essential to inter-compare the three reference systems using a common aircraft engine source  
9 to establish repeatability and intermediate precision of the sampling and measurement systems,  
10 and to estimate some of the uncertainties associated with the measurements of aircraft engine  
11 nvPM emissions.  
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19 In this paper, we present results of the Aviation-Particle Regulatory Instrumentation  
20 Demonstration Experiment (A-PRIDE) 5/ Studying, sAmpling and Measuring of aircraft  
21 ParticuLate Emissions (SAMPLE) III - SC03 campaign, the first multi-agency international  
22 collaborative project to inter-compare and evaluate the robustness of the Swiss, European, and  
23 North American standardized reference systems for the sampling and measurement of aircraft  
24 engine nvPM emissions. The measurements were performed from 28 July to 25 August 2013, on  
25 a leased CFM56-7B26/3 engine used during dedicated engine testing at the SR Technics engine  
26 test facility in Zürich, Switzerland.  
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## 34 35 **2 Methods**

### 36 37 *2.1 CFM56-7B26/3 engine*

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39 The engine used for the dedicated tests was a CFM56-7B26/3, which was leased for the  
40 duration of the campaign. The CFM56-7B engine is the sole powerplant for the Boeing 737NG  
41 family, and it is the most widely used engine in commercial aviation. The “/3” configuration  
42 (improved durability and emissions) was emissions-certified for gaseous pollutants and smoke in  
43 2006. The specific engine used during this campaign had accumulated 5009 flight hours during  
44 2000 flight cycles. The engine had representative operating and performance characteristics with  
45 minimal degradation and negligible oil consumption in the range of a brand-new engine. This  
46 engine was declared to be a representative reference engine to be used as a source of nvPM to  
47 inter-compare the standardized reference systems. Aircraft engine emissions certification-like tests  
48 for nvPM emissions were also performed with a single system (Durdina et al., 2017).  
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## 2.2 *Fuel*

The Jet A-1 fuel used for the emissions tests was provided by SR Technics. Seven fuel samples were collected during the measurement campaign, and the results of the fuel analysis performed by Intertek AG (Schlieren, Switzerland) for selected properties are presented in Table 1. The properties of the Jet A-1 fuel were all within the allowable range specified by ASTM D1655 (ASTM, 2019). The fuel also met all the specifications for emissions certification tests (ICAO, 2008) except that for naphthalenes content, which was below the lower limit requirement at the time (1%). This was also the case for a previous campaign conducted at SR Technics (Lobo et al., 2015a). The lower limit for naphthalenes specification was subsequently changed to 0% (ICAO, 2017). Overall, the different batches of Jet A-1 fuel used for the emissions tests had similar properties, thus eliminating fuel composition as a variable in the calculation of emission indices (EIs) and inter-comparison of the reference systems. An average hydrogen to carbon (H/C) ratio of 1.95 was used for the calculation of the number and mass emission indices for all three reference systems.

**Table 1: Properties of Jet A-1 fuel used during the A-PRIDE 5/SAMPLE III-SC03 campaign**

Property	Units	Method	Allowable Range	Test Fuel Samples <sup>†</sup>						
				29 Jul 2013	2 Aug 2013	5 Aug 2013	12 Aug 2013	17 Aug 2013	18 Aug 2013	25 Aug 2013
Density at 15°C	kg/m <sup>3</sup>	ASTM D4052	780-820	797.6	797.6	797.8	797.8	797.8	797.8	797.2
Kinematic viscosity at -20°C	mm <sup>2</sup> /s	ASTM D445	2.5-6.5	3.591	3.618	3.598	3.596	3.599	3.599	3.618
Distillation temperature	°C	ASTM D86	155-201	169	168	168	168	167	168	169
10% boiling point										
Final boiling point			235-285	265	265	265	261	263	264	264
Net heat of combustion	MJ/kg	ASTM D3338	42.86-43.50	43.3	43.3	43.3	43.3	43.3	43.3	43.3
Aromatics	volume %	ASTM D1319	15-23	17.7	17.7	17.4	18.0	17.7	17.7	17.5
Naphthalenes	volume %	ASTM D1840	0-3% <sup>‡</sup>	0.68	0.72	0.71	0.71	0.74	0.75	0.70

<sup>†</sup> All fuel samples were collected from the fuel line in the test cell except the on 5 Aug 13, which was collected directly from the fuel tanker

<sup>‡</sup> Original allowable range 1.0-3.5%, subsequently updated to 0-3%

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Smoke point	mm	ASTM D1322	20-28	21	21	21	21	22	21	22
Hydrogen content	mass %	ASTM D5291	13.4-14.3	14.18	14.18	14.28	14.04	13.96	14.00	13.76
Sulphur content	mass %	ASTM D5453	< 0.3%	0.053	0.033	0.039	0.039	0.042	0.042	0.042
H/C ratio (calculated)			1.84-1.99	1.97	1.97	1.99	1.95	1.93	1.94	1.90

## 2.4 Experimental Setup

A schematic of the experimental setup is shown in Fig. 1. Details of the various sections of the sampling and measurement systems are provided in the following sections.

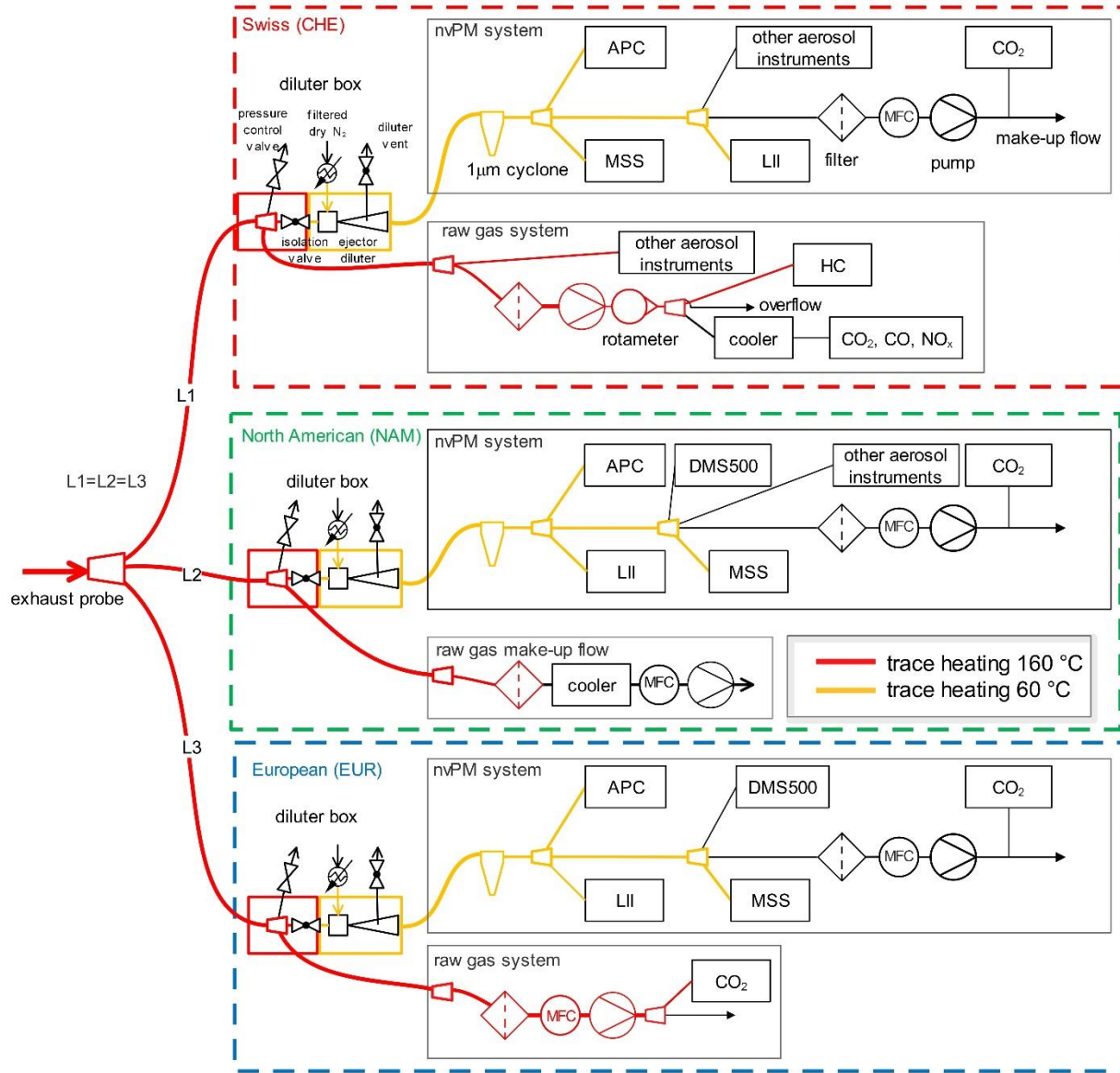


Figure 1: Experimental setup for CFM56-7B26/3 engine tests (not to scale)



#### 2.4.1 *Extractive sampling probes*

Two exhaust sampling probe assemblies were used to extract emission samples within 1 m of the engine exit plane. A fixed multi-point probe made of Inconel 625 alloy compliant with the requirements for nvPM sampling and measurements for emissions certification (ICAO, 2017) was used during a subset of the dedicated engine tests. It consisted of a cruciform with six sampling orifices located on each of the four arms, from which three orifices per arm were used to comply with the carbon balance check (air/fuel ratio estimated from the exhaust sample total carbon concentration agrees with the estimate based on engine air/fuel ratio). The sampling orifices were symmetrically located on circular radii from the center of the cruciform. Various configurations of the number and position of sampling orifices on the multi-point probe were evaluated during the campaign. The primary goal of this assessment was to ensure that the three reference systems operating in parallel were adequately supplied with sample flow, while still providing an exhaust representative sample and complying with the carbon balance check.

A traversable single-point probe used during previous campaigns at SR Technics (Lobo et al., 2015a) was used during piggy-back measurements and a subset of dedicated engine tests. The probe with an 8 mm ID orifice was made of Inconel 600 alloy. The probe's vertical traverse capability afforded representative sampling at the exit plane of different engine types operated in the test cell (for the CFM56-7B26/3 engine, the distance from the probe to the exit plane was 0.8 m). The specific vertical sampling location was optimized during testing, ensuring the carbon balance of the single-point probe maintained satisfactory agreement (< 10%) with the multi-point probe at all similar test conditions.

#### 2.4.2 *Sampling and measurement systems*

The probe was connected to a three-way splitter using a 5.2 m long, 8 mm ID thin-walled stainless steel tubing, electrically trace heated and insulated to maintain a temperature of  $160^{\circ}\text{C} \pm 15^{\circ}\text{C}$ . The exhaust samples were distributed to the three reference systems. Each reference system, while compliant with the specifications detailed in AIR6241, was independently assembled from non-identical components and evaluated prior to deployment in the campaign. Briefly, each system had a three-way splitter to distribute the exhaust sample to the line for raw gaseous emission measurements, the pressure control line to regulate the sample inlet pressure, and the nvPM line for nvPM mass- and number-based emission measurements (Fig. 1). Of the three reference

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4 systems, only the NAM system did not have measurements of raw gaseous emissions on the raw  
5 exhaust line. Instead, a pump was used to draw flow down the gaseous emissions line in order to  
6 achieve a prescribed flow velocity in each leg of the splitter, and to be consistent with the CHE  
7 and EUR system flowrates. The undiluted CO<sub>2</sub> concentration data from the CHE reference system  
8 were used to calculate dilution factors for all three systems. The variability in undiluted CO<sub>2</sub>  
9 concentrations measured by the CHE and the EUR reference systems was evaluated. The slope of  
10 a linear interpolation comparing the undiluted CO<sub>2</sub> concentrations in the EUR system to those  
11 from the CHE system was 0.993 (R<sup>2</sup>=0.997). This variability is within the uncertainty of the CO<sub>2</sub>  
12 measurements, estimated to be less than 4% across all test points.  
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21 The exhaust sample for the nvPM line was diluted with particle-free synthetic air (purity  
22 99.999%) using a Dekati DI-1000 ejector diluter. The dilution factors achieved using the DI-1000  
23 ejector diluter are highly dependent on the sample inlet pressure and diluent air pressure  
24 (Giechaskiel, Ntziachristos, & Samaras, 2004). The dilution factor at nominal sample and diluent  
25 inlet pressures was 8-14 (as specified in the SARPs); however, the range extended from 7 to 20  
26 for some of the dilution factor sensitivity tests performed during the campaign. The diluted sample  
27 was transferred to the real-time diagnostic instruments by a 24 - 25 m long, carbon-loaded,  
28 electrically grounded polytetrafluoroethylene (cPTFE) line maintained at a temperature of 60°C ±  
29 15°C followed by a sharp cut cyclone with a 1 µm cut size. The sample lines for each reference  
30 system had slightly different internal diameters, but well within the range 7.59 - 8.15 mm as  
31 specified by AIR6241. The diluted exhaust sample flow rate in the cPTFE line was maintained at  
32 25 slpm ± 2 slpm. Another three-way splitter distributed the particle-laden flow to the nvPM  
33 number instrument, [the nvPM mass instruments](#), and the excess flow line for CO<sub>2</sub> concentration,  
34 and other ancillary measurements that were not a requirement of the standardized system.  
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46 All three reference systems measured nvPM number emissions using an AVL Particle  
47 Counter (APC) Advanced (Giechaskiel et al., 2010), which consists of a primary dilution stage  
48 with a rotating disk diluter, a catalytic stripper with a sulphur trap (volatile particle remover, VPR)  
49 maintained at 350°C, a secondary dilution stage with a porous tube diluter, and an n-butanol-based  
50 condensation particle counter (CPC; TSI 3790E), with a 50% cut-off diameter ≈10 nm and 90%  
51 count efficiency at ≈15 nm (Lobo et al., 2015a). All CPC's were operated in single count mode by  
52 increasing the APC dilution factor when CPC concentrations approached 10,000 particles/cm<sup>3</sup>.  
53 The diluent used for the APC was synthetic air. The nvPM mass emissions were measured using  
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4 both an Artium Laser Induced Incandescence LII 300 (LII) (Snelling et al., 2005) and an AVL  
5 Micro Soot Sensor (MSS) (Schindler et al., 2004) on all three reference systems. Additionally, two  
6 Cambustion DMS500 fast mobility spectrometers (Reavell et al. 2002) were installed on the excess  
7 flow/ancillary line, one each on the EUR and NAM reference systems to measure PM size  
8 distributions. A compact time of flight aerosol mass spectrometer (CToF-AMS) (Drewnick et al.  
9 2005) to obtain chemical composition information, and a Cavity Attenuated Phase Shift (CAPS)  
10 PM extinction monitor (Yu et al., 2011) to measure nvPM mass were installed on the ancillary line  
11 on the NAM reference system. Further aerosol instrumentation was also deployed in the CHE  
12 reference system to measure particle effective density (Durdina et al., 2014) and chemical  
13 composition (Abegglen et al., 2016).  
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### 24 2.4.3 Instrument calibrations

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26 The LII 300 and MSS instruments were calibrated to the NIOSH 5040 protocol (NIOSH,  
27 2003) for EC using thermal optical analysis (TOA) in accordance with AIR6241 in two batches at  
28 NRC-Metrology, Ottawa, Canada one month prior to the start of the campaign. The NRC inverted-  
29 flame burner (Coderre et al., 2011) was used as the source of black carbon (BC) particles. The BC  
30 particles generated by the inverted-flame burner were diluted using filtered air, and the diluted  
31 sample was divided using a splitter and then directed to two 1 $\mu$ m cyclones – one was upstream of  
32 the dual-stage filter collection system, and the other upstream of the nvPM mass instruments to be  
33 calibrated. Equal length (2 m) heated cPTFE 3/8" OD tubing was used from splitters to all  
34 measurement devices. The entire sampling system was heated to 60°C. The MSSs from all three  
35 reference systems and the LII 300 from the NAM reference system were calibrated in the first  
36 batch with 6 or more repeats at target mass concentrations of 0, 50, 100, 500, and 1000  $\mu$ g/m<sup>3</sup>. The  
37 remaining two LII 300s were calibrated in the second batch with 3 or more repeats at target mass  
38 concentrations of 0, 100, 250, and 500  $\mu$ g/m<sup>3</sup>. The correlation of mass concentration measured by  
39 the instruments to NIOSH 5040 EC was >0.995 in all cases.  
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51 The APC for all three reference systems had its annual calibration and maintenance  
52 performed by the manufacturer within nine months of the campaign. Combustion soot from a  
53 miniature combustion aerosol standard (miniCAST) generator was used as the source to establish  
54 penetration through the VPR. CO<sub>2</sub> calibration gas was used to verify dilution factors for the two-  
55 stage dilution of the APC. As part of the calibration procedure, VPR performance was evaluated  
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4 in terms of volatile particle removal efficiency. The CPC linearity and counting efficiency were  
5 also determined and adjusted (counting efficiency at 10 nm > 50%). All three reference systems  
6 had similar particle penetration profiles for the VPR, and volatile particle removal efficiencies of  
7 99.99%. The CHE and NAM reference systems had similar CPC counting efficiencies of 76% and  
8 92% at 10 nm and 15 nm, respectively, while the EUR reference system had a much lower CPC  
9 counting efficiency of 53% at 10nm and slightly higher CPC counting efficiency of 98% at 15 nm.  
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## 18 **2.5 Test Matrix**

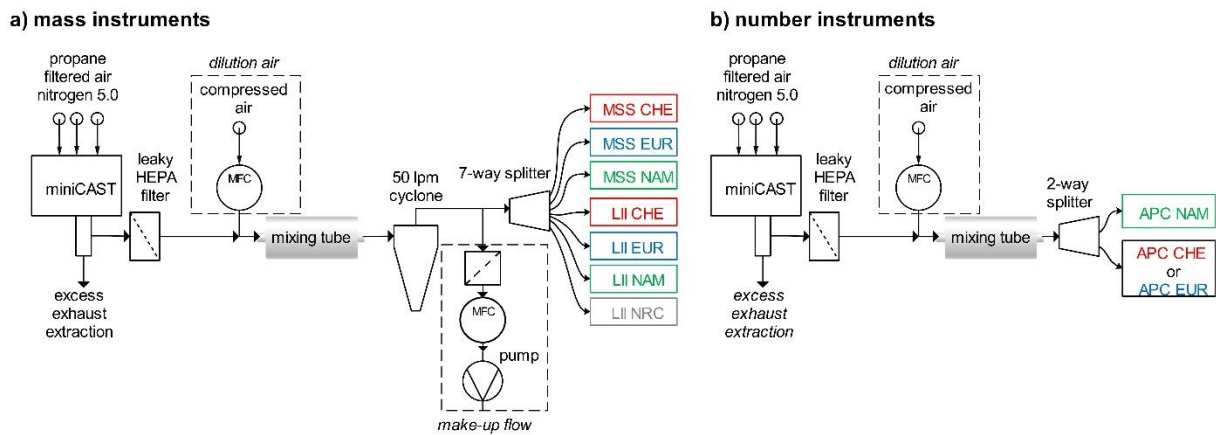
19 The dedicated engine tests with the CFM56-7B26/3 engine started with a warm-up  
20 sequence used for conditioning the probe and the sampling systems. The warm-up sequence  
21 consisted of running the engine at five test points from ground idle to 85% sea-level static thrust  
22 for durations of 5 minutes each. The parameter used for setting the engine thrust was the combustor  
23 inlet temperature, T3. The T3 values were based on a correlation of sea-level static thrust with T3  
24 corrected to ISA conditions (15°C, 1 atm). The full test matrix following the warm-up sequence  
25 consisted of 12 points on a descending power curve, starting at maximum continuous thrust (which  
26 was limited by the ambient conditions during the test) and ending at idle. These test points included  
27 the four thrust settings corresponding to the LTO cycle as well as an additional point at 65%. The  
28 duration of each test point was 10 minutes. A subset of the full test matrix was run during most  
29 dedicated engine tests. Table 2 lists the inter-comparison experiments performed during the  
30 measurement campaign, along with the range of ambient conditions recorded during each test.  
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**Table 2: Overview of experiments performed for Reference Systems inter-comparison**

Date	Ambient Condition Ranges			Engine/Source	Description	Sampling Probe	Reference System		
	T (° C)	P (kPa)	RH (%)				Swiss (CHE)	European (EUR)	North American (NAM)
28 Jul 2013	28.5-34.0	96.0-96.2	24-42	CFM56-7B26/3	AFR check (probe tips 1, 4, 6)	Multi-point	✓	✓	✗
29 Jul 2013	17.3-17.8	96.6-96.8	87-95	CFM56-7B26/3	AFR check (all 24 probe tips)	Multi-point	✓	✓	✗
2 Aug 2013				miniCAST 5201C	Instrument comparison	-	✓	✓	✓
3 Aug 2013	27.4-30.8	96.9-97.0	28-39	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✓	✓
4 Aug 2013	18.0-26.3	97.1-97.5	45-82	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✓	✓
5 Aug 2013	24.2-25.6	96.7-96.8	54-57	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✓	✓
10 Aug 2013	21.1-22.5	97.1-97.2	44-53	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✗	✓
11 Aug 2013	18.4-24.2	96.9-97.1	39-56	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✗	✓
12 Aug 2013	20.4-26.1	96.6-96.8	34-50	CFM56-7B26/3	Reference Systems comparison	Single-point	✓	✗	✓
24 Aug 2013	17.0-20.2	96.3-96.5	68-92	CFM56-7B26/3	Reference Systems comparison	Multi-point	✓	✓	✗

## 2.6 miniCAST

A miniCAST soot generator 5201C (Jing Ltd) was used as a surrogate emissions source to compare the nvPM mass and number diagnostic instruments used in the three reference systems prior to the dedicated engine tests. The miniCAST was operated at the following flowrates - propane: 0.06 lpm, N<sub>2</sub> mixing gas: 0 lpm, oxidation air: 1.55 lpm, N<sub>2</sub> quench air: 7 lpm, dilution air: 20 lpm, such that a high elemental carbon (EC) fraction (>80%), determined from TOA, was produced in the exhaust stream (Durdina et al., 2016). The mean size of particles at the miniCAST setting selected was ~130 nm. This test was performed to verify that the operation and performance of the instruments were optimal after transport to the SR Technics engine test facility. The configurations for the nvPM mass and number instrument comparisons using the miniCAST are presented in Figure 2.



**Figure 2: Configuration for the mass (a) and number (b) instrument comparisons using the miniCAST 5201C soot generator**

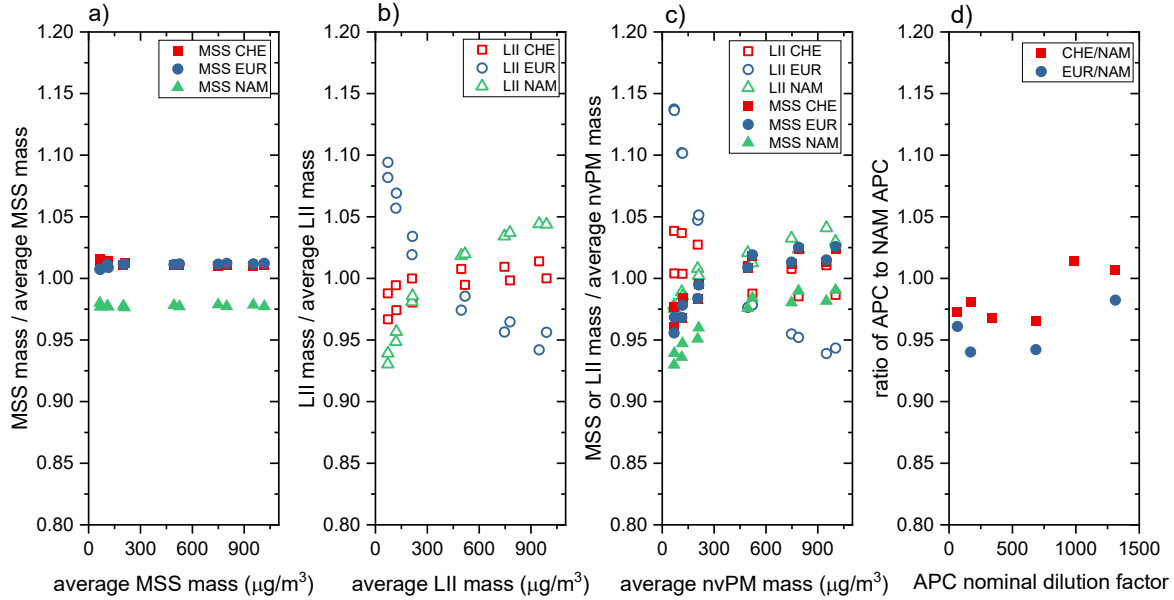
For the nvPM mass instrument comparison, the miniCAST exhaust was passed through a HEPA filter dilution bridge and then diluted with compressed air to achieve target mass concentrations of 50, 100, 250, 500, 750, and 1000  $\mu\text{g}/\text{m}^3$ . The diluted exhaust sample was then transferred through a mixing tube, a cyclone with a 1  $\mu\text{m}$  cut-point at 50 lpm, and a 7-way splitter to the instruments under test – 6 instruments, (3 MSSs and 3 LII 300s) two from each reference system, and another LII 300 which was used to monitor source concentration levels during the course of the instrument comparisons. Only data from the instruments in the reference systems were used in the analysis. The calibration factors determined during the NRC-Metrology

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4 calibrations from the NIOSH 5040 EC protocol for all six nvPM mass instruments were applied  
5 prior to the comparison study. For the nvPM number instrument comparison, the miniCAST  
6 exhaust was diluted with compressed air and split into two legs – one providing a sample to the  
7 APC from the NAM reference system, and the other to the APC from either the CHE or EUR  
8 reference systems.  
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### 14 **3 Results and Discussion**

#### 15 *3.1 Instrument comparisons using the miniCAST*

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20 The results of the nvPM mass and number instrument comparisons conducted using the  
21 miniCAST as a source of nvPM are presented in Figure 3. The nvPM mass concentrations as  
22 measured by the MSS (Fig 3a) and LII 300 (Fig 3b) for each reference system were averaged for  
23 60 seconds, and each instrument was then compared against the ensemble average. The data from  
24 the MSS and LII 300 instruments were tightly bound with a relative standard deviation (RSD) of  
25 1.95% and 4.5%, respectively. A relatively higher degree of scatter ( $\pm 10\%$ ) was observed with the  
26 LII 300 data for nvPM mass concentrations below  $100 \mu\text{g}/\text{m}^3$ . Overall, the six nvPM mass  
27 instruments were within an RSD of 4.1% over the range of target nvPM mass concentrations  
28 explored. An RSD of 2.1% was observed between the 3 APCs used for the nvPM number  
29 measurements when the CHE and EUR reference system APCs were compared against the NAM  
30 reference system APC (Fig 3d). The repeatability of the miniCAST as reported by the  
31 manufacturer is  $\pm 5\%$ . Larger differences have been observed from inter-day experiments (Moore  
32 et al. 2014), however, for the instrument comparison test the miniCAST settings were stable with  
33 RSDs in the mean nvPM mass and number concentrations  $< 5\%$ . Since the differences observed  
34 between the instruments were of a similar magnitude, the nvPM mass instruments and the nvPM  
35 number instruments were assessed as being in statistical agreement on the miniCAST source.  
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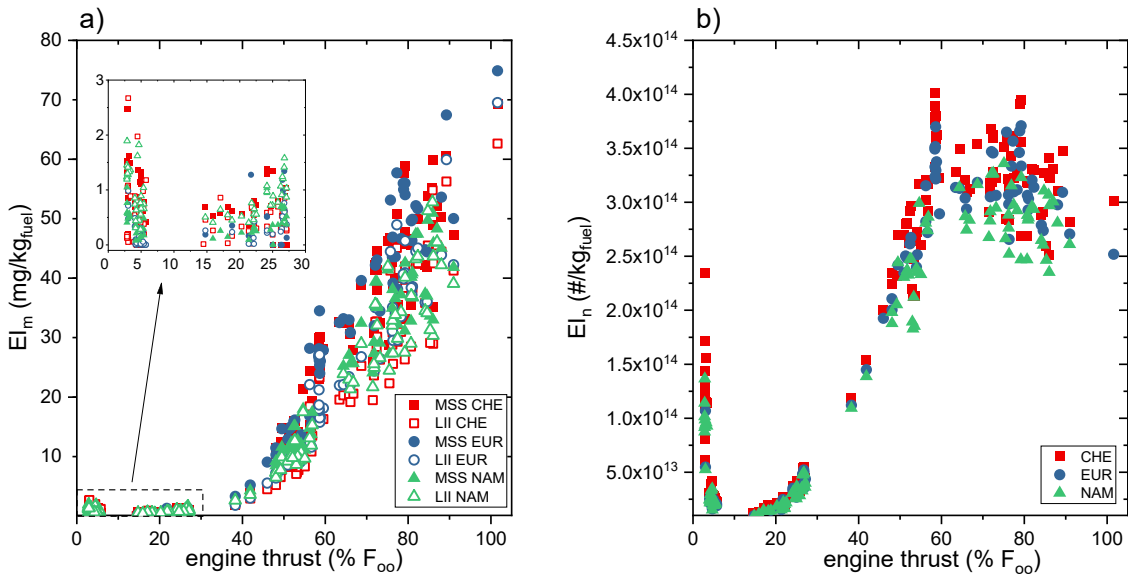
**Figure 3: Comparison of nvPM mass (a, b, c) and number (d) instruments using the miniCAST 5201C soot generator**

### 3.2 *nvPM mass and number emission profiles for the CFM56-7B26/3 engine*

The nvPM mass and number concentrations were converted to nvPM mass-based emission index ( $EI_m$ ) and nvPM number-based emission index ( $EI_n$ ), respectively, using the measured nvPM and gaseous emissions concentration and following the procedures specified in AIR6241 (SAE, 2013). The EIs are reported at a standard temperature of 273.15 K and standard pressure of 101.325 kPa. The nvPM mass and number emission indices were not corrected for either the thermophoretic loss in the sample extraction system or for size-dependent diffusional and inertial losses that occurred in the sampling and measurement systems, and CPC efficiencies for nvPM number. The nvPM mass and number emission indices for the CFM56-7B26/3 engine as a function of percent rated thrust are presented in Figure 4. For the nvPM mass emission index profile, the emissions for this engine were generally higher at idle conditions (3-7% rated thrust), decreased to a minimum at low engine thrust conditions (15-30% rated thrust), and increased linearly to maximum rated thrust. The nvPM number emission index also exhibited behavior similar to that of nvPM mass emissions at idle and low engine thrust conditions. However, the emissions increased up to a maximum at ~60% rated thrust and then slightly decreased up to the maximum rated thrust. The nvPM mass and number emissions profiles shown here are consistent with previously reported profiles for a CFM56-7B24/3 engine (which is the same engine model but



rated at a lower take-off thrust) (Lobo et al., 2015a). ~~The theoretical total uncertainties in the nvPM EIs presented are estimated to be ~22% for EI mass and ~25% for EI number. These estimates are based on typical uncertainty values for the nvPM mass and number instruments (including calibration uncertainty), CO<sub>2</sub> measurement, and the determination of the dilution factor in the nvPM number instrument (SAE, 2013). These estimates do not account for either the particle losses in the sampling and measurement systems or the increased uncertainty for nvPM mass measurements near the limit of detection (LOD).~~



**Figure 4: nvPM mass (a) and number (b) emission index profiles for the CFM56-7B26/3 engine**

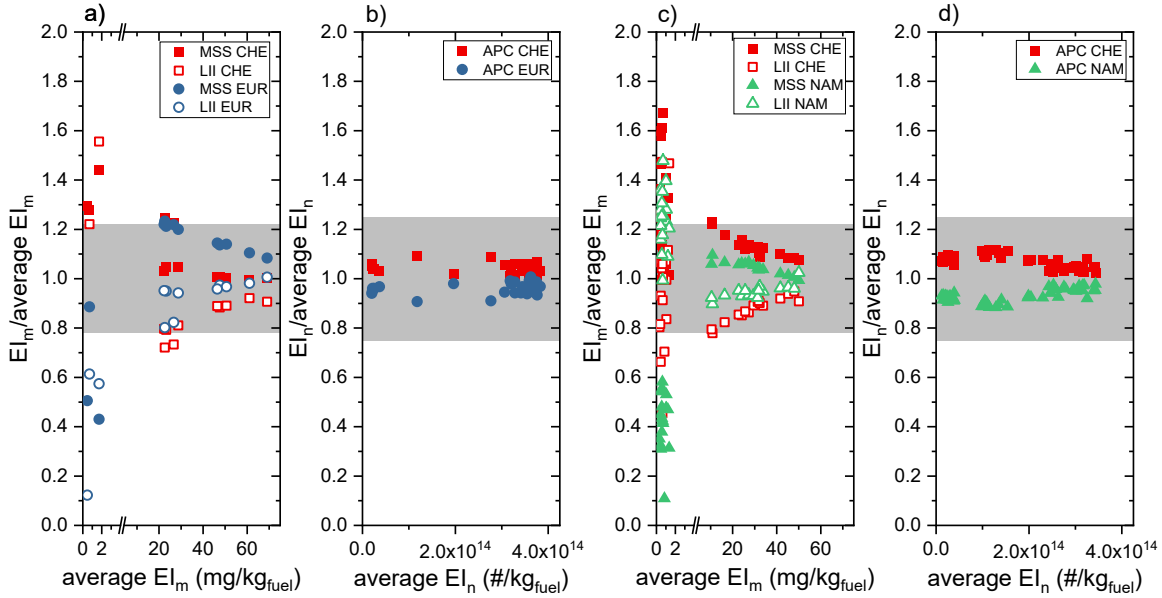
### 3.3 2-way reference system comparisons (CHE-EUR; CHE-NAM)

Comparison of nvPM mass and number emissions for pairs of reference systems, i.e., CHE-EUR and CHE-NAM was performed during the test campaign and is presented in Figure 5. The multi-point sampling probe was used for the CHE-EUR comparison, while the single-point probe was used for the CHE-NAM comparison. However, since the extracted exhaust samples were representative of the engine emissions, the use of a particular sampling probe did not influence the comparisons between the reference systems. The 2-way comparisons with reference systems in parallel were performed to assess system to system differences without the additional complexity of including a third system (also performed and described in the next section). The nvPM mass

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4 and number emission index for each instrument in the reference system pairs were averaged and  
5 compared against the ensemble average for a specific pair-wise comparison. The average nvPM  
6 mass and number emission indices for the CHE-EUR comparison were slightly higher than the  
7 CHE-NAM comparison because of the higher engine thrust conditions achieved during the  
8 respective tests.  
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13 The theoretical total uncertainties in the nvPM EIs presented are estimated to be ~22% for  
14 EI mass and ~25% for EI number. These estimates are based on typical uncertainty values for the  
15 nvPM mass and number instruments (including calibration uncertainty), CO<sub>2</sub> measurement, and  
16 the determination of the dilution factor in the nvPM number instrument (SAE, 2013). These  
17 estimates do not account for either the particle losses in the sampling and measurement systems or  
18 the increased uncertainty for nvPM mass measurements near the limit of detection (LOD).  
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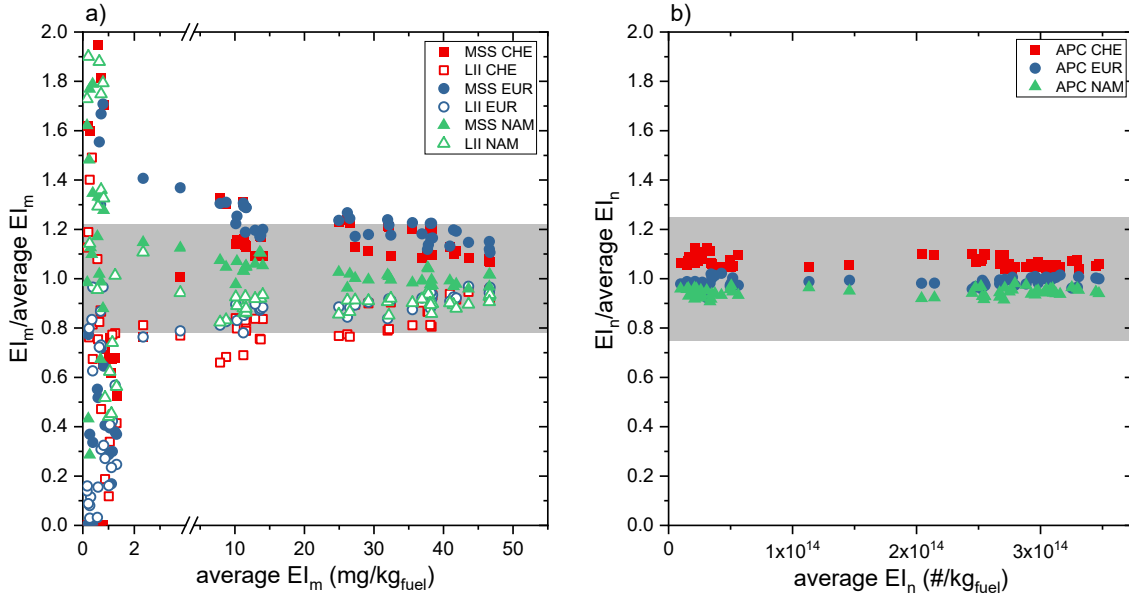
28 All instruments in the three reference systems were generally  $\pm 22\%$  of the average nvPM  
29 mass emission index and  $\pm 25\%$  of the average nvPM number emission index (grey shaded area  
30 in Fig. 5), except for the nvPM mass emission index which exhibited a ~~20~~22% variation at low  
31 nvPM mass concentrations levels (corresponding to EIs of  $< 30$  mg/kg), and significantly higher  
32 differences as the instruments approached the LOD ( $3 \mu\text{g}/\text{m}^3$ , corresponding to EIs of  $< 2$  mg/kg).  
33 The high variability in mass at low concentrations was exhibited for both types of nvPM mass  
34 instruments (MSS and LII 300). This trend for high variability in nvPM mass at low concentrations  
35 is consistent with results for emissions measurements of other engine types (Lobo et al., 2015a;  
36 Lobo et al., 2016). For nvPM number, the CHE reference system registered consistently higher  
37 values compared to the EUR and NAM reference systems. The overall magnitude of variation for  
38 each instrument from the average was consistent for the two pairs of reference system  
39 comparisons, i.e  $\pm 15\%$ . It should be noted that since only a single type of nvPM number  
40 instrument was compared, there is no information on the uncertainty associated with using  
41 different types of nvPM number instruments that meet the specifications.  
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**Figure 5: Comparison between CHE and EUR reference systems (a, b) and CHE and NAM reference systems (c, d) for nvPM mass and number emission indices**

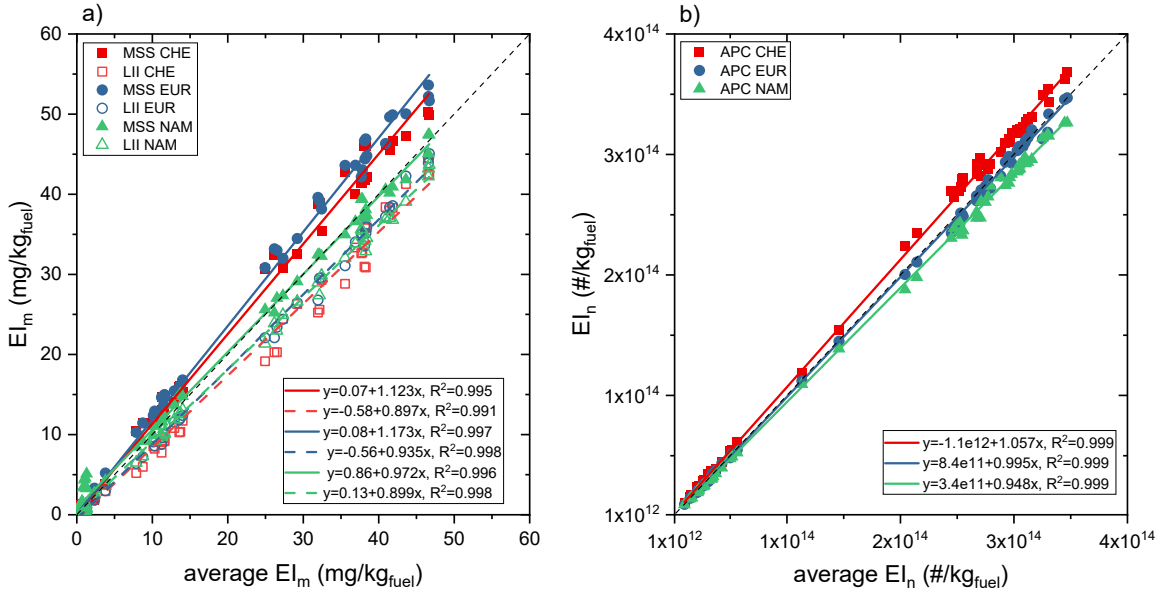
### 3.4 3-way reference system comparisons (CHE-EUR-NAM)

All three reference systems were compared simultaneously using the CFM56-7B26/3 engine. The comparisons between the reference systems in terms of the ratio of the nvPM mass and number emission index to the average emission index as a function of the average emission index are presented in Figure 6. Similar trends in nvPM mass and number EIs during the 2-way comparisons are also observed for the 3-way comparisons. The average nvPM mass and number EIs for the 3-way reference system comparisons are lower than those for the 2-way comparisons because the nvPM emissions produced by the engine for these tests were lower. Variability in nvPM mass emission index was higher than 20% for EIs up to 40 mg/kg fuel (corresponding to a mass concentration of  $\sim 95 \mu\text{g}/\text{m}^3$  at the instrument), while the variability in nvPM number emission index was  $\pm 15\%$  for all test conditions.



**Figure 6: Comparison between CHE, EUR, and NAM reference systems for nvPM mass (a) and number emission index (b).**

It is also informative to view the nvPM mass and number emission indices for the instruments used in the three reference systems using parity plots as shown in Figure 7. The EI data reported for each instrument are plotted against the average EI. While these plots are not suitable to illustrate differences at low concentration levels, they provide an overall magnitude of variability between the instruments. As can be seen in Figure 7, the nvPM  $EI_m$  and  $EI_n$  for each instrument ~~was~~ were well correlated with the average. The nvPM  $EI_n$  for all three systems was within  $\pm 6\%$  of the average. For the nvPM  $EI_m$ , the magnitude of the differences was  $\sim 10\%$  for the LII 300 and  $\sim 15\%$  for the MSS. Ideally, quartz filters would have been collected in parallel with the real-time instruments to determine EC content using TOA, and then used as the reference to compare the nvPM mass instruments. The filter collection for TOA was not performed during the campaign due to limitations on sampling time at each test condition preventing adequate sample to be collected for analysis.



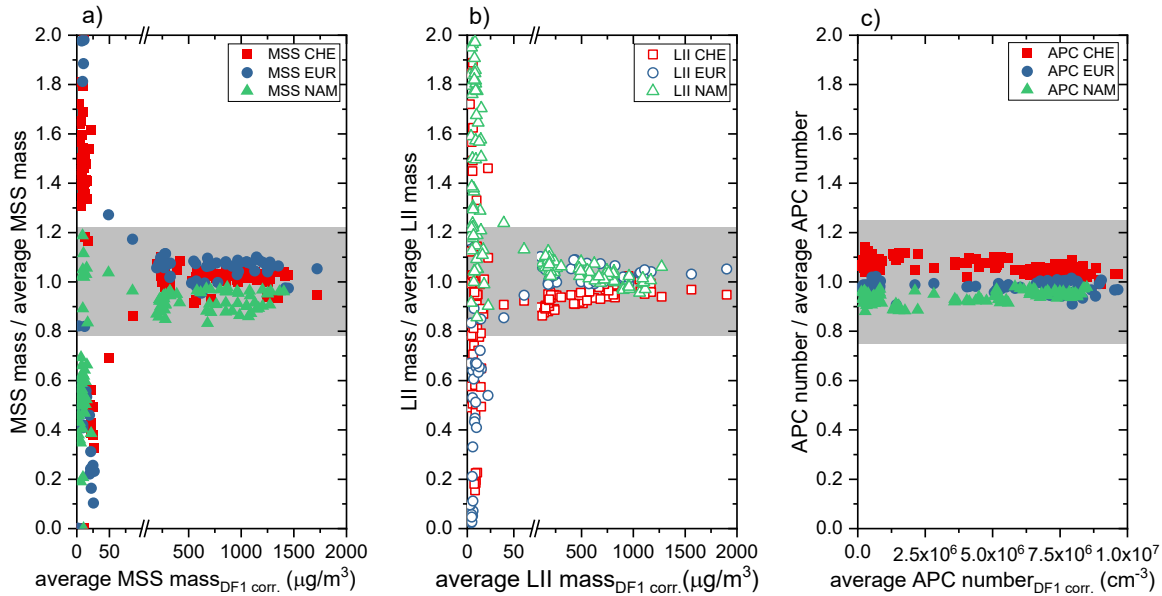
**Figure 7: Parity plot comparisons of CHE, EUR, and NAM reference systems for nvPM mass (a) and number emission index (b).**

### 3.5 Reference system comparisons for nvPM mass and number concentration

The comparison between the reference systems in terms of nvPM mass and number concentrations for each type of measurement instrument is presented in Figure 8 to assess performance on a concentration basis (the primary output of the instrument). The concentration data at the measurement location (Figure 8) have been corrected for dilution since each system had slightly different dilution factors. All data recorded during the campaign when at least 2 systems were operating in parallel are included in this analysis. The comparison between the three reference systems is presented as a function of the ensemble averages for the different types of measurement instruments, i.e. MSS and LII 300 for nvPM mass, and APC for nvPM number.

As was previously reported, the largest differences between the three reference systems for nvPM mass were observed for dilution corrected nvPM mass concentrations  $< 50 \mu\text{g}/\text{m}^3$  at the measurement location ( $\sim 5 \mu\text{g}/\text{m}^3$  at the instrument). Beyond this threshold, the particular nvPM mass instrument type, i.e. MSS or LII 300, in the reference systems were within 20% of the instrument-specific average mass concentrations. For a given nvPM mass instrument type, the variability in the measured nvPM mass emissions is constrained in a narrow range, which is not the case when both nvPM mass instrument types are included in the analysis (see Figures 5 and

6). Unlike the instrument comparisons with the miniCAST (Figures 3 a and b), both types of nvPM mass instruments each demonstrate similar variability, exceeding 20% only below  $50 \mu\text{g}/\text{m}^3$  (Figures 8 a and b) on engine exhaust. For nvPM number, all three reference systems were well within 20% of the dilution corrected average concentration over the entire range of values recorded.



**Figure 8: Comparison between CHE, EUR, and NAM reference systems for nvPM mass (a,b) and number concentrations (c). The concentrations are reported at the measurement location and corrected for dilution.**

### 3.6 Variability

The variability in nvPM mass and nvPM number emissions was computed by calculating the RSD of the ratio of the respective EI to the average EI (data from Figures 5 and 6). This method for determining variability was adopted to focus on the relative response of the instruments (as a function of concentration), and to decouple the thrust dependency of the EIs and variability in ambient temperature. The measurement campaign was conducted over the course of a month, and the wide range of ambient conditions affected the nvPM emissions produced by the engine.

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4 Previous studies have also reported on the impact of ambient conditions such as temperature on  
5 nvPM emissions variability (Lobo et al., 2015a).  
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8 The three reference systems were inter-compared to establish repeatability and  
9 intermediate precision of the sampling and measurement systems. Repeatability is defined as the  
10 variability of many measurements where the same equipment and operator are used to make  
11 repeated measurements over a short time period, while intermediate precision refers to the  
12 variability of measurements when only some of the four precision conditions (time, calibration,  
13 equipment, operator) are different (JCGM, 2012). The variability in nvPM mass emissions for  
14 repeatability (intra-system) and intermediate precision (inter-system) comparisons are presented  
15 in Table 3 as a function of average nvPM concentrations, with lowest concentrations of nvPM  
16 mass and number grouped in the case of the CFM56-7B26/3 engine at low engine thrust ranges(3-  
17 30%), with increasing concentrations averaged at medium engine thrust (38-60%), and high engine  
18 thrust (63-101%). A similar analysis for nvPM number emissions for intermediate precision  
19 comparisons is presented in Table 4. Intra-system (repeatability) variability for nvPM number  
20 emissions is not considered since each reference system used the same instrument type (APC) for  
21 the measurement.  
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33 The variability for nvPM mass was highest in the lowest mass concentrations (low engine  
34 thrust) range, where the average instrument concentration was below the LOD for both types of  
35 nvPM mass instruments at  $0.7 \mu\text{g}/\text{m}^3$ . The resolution of the MSS instruments used during the  
36 campaign was  $1 \mu\text{g}/\text{m}^3$ , whereas the LII 300 had a resolution of  $0.01 \mu\text{g}/\text{m}^3$ . The higher variability  
37 of the MSS compared to LII 300 at the LOD is likely introduced through the resolution of the  
38 instrument. For medium and high concentrations (successively higher engine thrust ranges), the  
39 variability within a system and between the reference systems was  $<13\%$ . For nvPM number, the  
40 variability was  $<3\%$  across the engine thrust ranges. The sensitivity of nvPM number to the limit  
41 of detection was not a factor since the measured concentration was significantly above the LOD.  
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**Table 3: Variability in nvPM mass emissions**

Comparison	Average instrument mass concentration	0.7 µg/m <sup>3</sup>						28 µg/m <sup>3</sup>						87.8 µg/m <sup>3</sup>					
	Thrust range	Low (3-30%)						Medium (38-60%)						High (63-101%)					
	Reference System	CHE MSS	EUR MSS	NAM MSS	CHE LII	EUR LII	NAM LII	CHE MSS	EUR MSS	NAM MSS	CHE LII	EUR LII	NAM LII	CHE MSS	EUR MSS	NAM MSS	CHE LII	EUR LII	NAM LII
Repeatability (Intra-system)	CHE	62%			48%			6.9%			10%			5.5%			7.1 %		
	EUR		65%			71%			3.8%			5.7%			3.7%			4.8%	
	NAM			60%			33%			3.1%			3.7%			2.4%			2.6%
Intermediate precision (Inter-system)	CHE-EUR and CHE-NAM	16%	40%	36%	32%	62%	30%	8.7%	1.2%	1.8%	4.6%	8%	2%	5.8%	4.1%	2.2%	5.1%	6.9%	2.7%
	CHE-EUR-NAM	126%	109%	68%	71%	75%	56%	12.8%	5.2%	5.5%	8%	5.3%	7.9%	5.2%	3.2%	5.6%	8%	3.7%	2.4%

**Table 4: Variability in nvPM number emissions**

Comparison	Average instrument number concentration (corrected for 2-stage dilution in the APC)	4.9×10 <sup>4</sup> /cm <sup>3</sup>			5.4×10 <sup>5</sup> /cm <sup>3</sup>			7.6×10 <sup>5</sup> /cm <sup>3</sup>		
	Thrust range	Low (3-30%)			Medium (38-60%)			High (63-101%)		
	Reference System	CHE	EUR	NAM	CHE	EUR	NAM	CHE	EUR	NAM
Intermediate precision (Inter-system)	CHE-EUR and CHE-NAM	2.1%	2.3%	1.8%	2.2%	2.2%	1.7%	1.8%	3.1%	1.0%
	CHE-EUR-NAM	2.4%	2.7%	1.7%	2.0%	1.1%	2.4%	1.0%	1.3%	0.9%



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4 The other significant contributor to the variability in nvPM EI<sub>m</sub> and nvPM EI<sub>n</sub> is the CO<sub>2</sub>  
5 concentration. The diluted CO<sub>2</sub> measurements were used to calculate the EIs from mass and  
6 number concentrations measured by each reference system. It was not possible to evaluate the  
7 variability of the diluted CO<sub>2</sub> measurements. Each reference system had slightly different  
8 dimensions for the ejector-diluter vent, which resulted in subtle differences in overall dilution  
9 factors. A comparison of the CO<sub>2</sub> analyzers measuring the same exhaust sample on the diluted  
10 nvPM line during the engine tests was not performed. However, the variability in undiluted CO<sub>2</sub>  
11 concentrations measured by the CHE and the EUR reference systems was evaluated. The slope of  
12 a linear interpolation comparing the -undiluted CO<sub>2</sub> concentrations in the EUR system to those  
13 from the CHE system was 0.993 (R<sup>2</sup>=0.997).  
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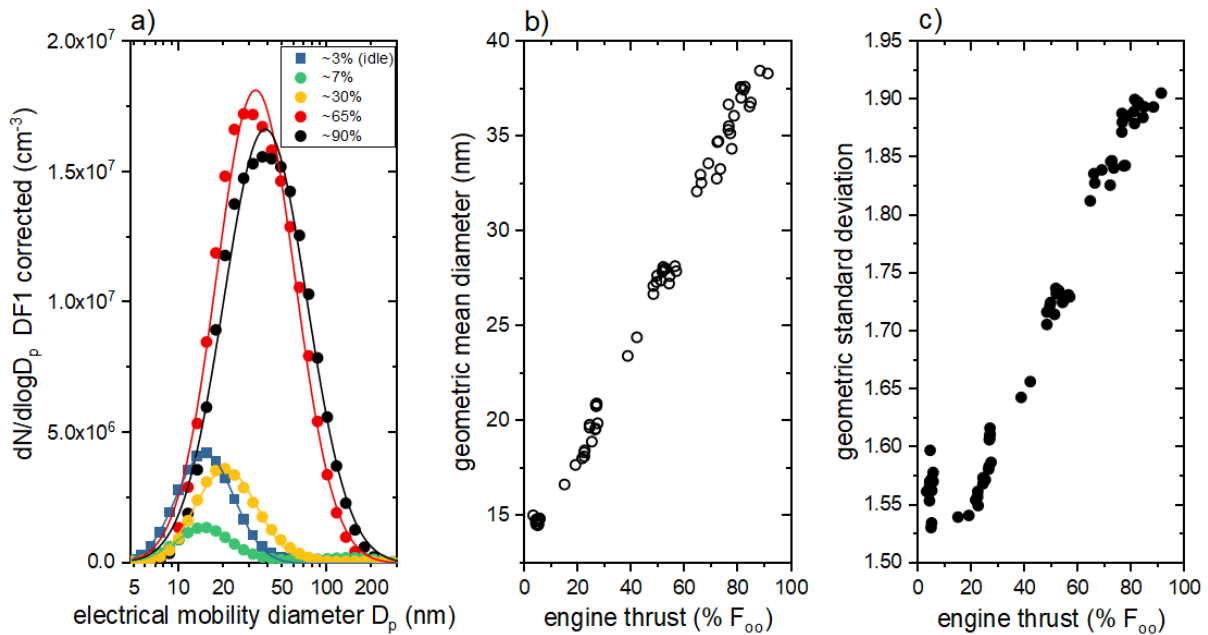
22 All three reference systems were built to and compliant with the specifications for the  
23 standardized system detailed in AIR6241, and in this case used nominally identical nvPM  
24 instruments. Hence, the differences in particle losses in these three sampling and measurement  
25 systems are expected to be negligible compared to the variability in other factors described  
26 previously.  
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32 Building on the knowledge gained from this campaign, several changes to the instrument  
33 performance and calibration protocols were implemented. The resolution of the MSS instruments  
34 was updated to 0.01 µg/m<sup>3</sup>. The procedure to demonstrate conformity of the nvPM mass  
35 instruments to performance specifications was updated to include an additional applicability  
36 criterion for validation of the calibration to EC on aircraft turbine engine exhaust. The limit of  
37 detection of the nvPM mass instruments was also lowered from 3 µg/m<sup>3</sup> to 1 µg/m<sup>3</sup> (ICAO, 2017).  
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### 44 **3.7 Size distributions**

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46 The standardized protocol for aircraft engine nvPM mass and number emissions does not  
47 specify a measurement of particle size distribution. However, a size distribution measurement is  
48 being considered for future standardized methodologies for particle loss correction. Since particle  
49 loss mechanisms such as diffusion and inertial losses are size-dependent, measurement of size  
50 distributions along with nvPM number and mass concentration provides information to estimate  
51 particle loss factors. These loss factors can then be used to calculate nvPM emissions at the engine  
52 exit plane. Engine exit plane emissions would be more relevant for aircraft engine nvPM emissions  
53 inventory and impact assessments. The size distributions for the CFM56-7B26/3 engine along with  
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characteristic parameters – geometric mean diameter (GMD) and geometric standard deviation (GSD) are presented in Figure 9 as a function of engine thrust setting. These size distributions were obtained with the DMS500 installed on the ancillary line of the EUR reference system and corrected for primary dilution (DF1) in the ejector diluter. The size distributions could be approximated to lognormal distributions ( $R^2 > 0.97$ ) with GMD ranging from 15nm at idle to 38nm at 90% rated thrust, and GSD varying between 1.53 and 1.92. The magnitude and general increasing trend of GMD and GSD with engine thrust setting are consistent with previously reported values for this engine type (Lobo et al., 2011; Lobo et al., 2015a; Durdina et al., 2017; Elser et al., 2019).



**Figure 9: Particle size distributions (a) and characteristic parameters – GMD (b) and GSD (c) for the CFM56-7B26/3 engine**

#### 4 Conclusions

Three reference systems for aircraft engine nvPM emissions measurement – the Swiss (CHE) system, the European (EUR) system, and the North American (NAM) system – were developed in compliance with the specifications for the standardized sampling and measurement methodology. The first and only inter-comparison to date of these three reference systems was performed at the SR Technics engine test facility in Zürich, Switzerland using a commercial CFM56-7B26/3 aircraft engine as the emissions source to establish repeatability and intermediate

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4 precision of the sampling and measurement systems. All three reference systems measured nvPM  
5 number concentration using an APC, and nvPM mass concentration was measured using both an  
6 LII 300 and an MSS. The nvPM mass and number concentrations were converted to their  
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8 respective emission indices for comparison. The specifications for the standardized sampling and  
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10 measurement system implemented in the three reference systems were robust, as demonstrated by  
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12 the variability observed between the systems. During the dedicated engine tests with the CFM56-  
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14 7B26/3 engine, all instruments in the three reference systems were generally within 30% of the  
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16 average nvPM mass emission index (determined with different nvPM mass instrument types and  
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18 manufacturers) and 15% of the average nvPM number emission index (determined with the same  
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20 nvPM number instrument type and manufacturer) (see Fig. 6). The only exception was for the  
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22 mass instruments, which exhibited a higher variation as the concentration levels approached the  
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24 LOD of 3  $\mu\text{g}/\text{m}^3$ . A comparison between the three reference systems as a function of the  
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26 measurement instrument type revealed that similar measurement methodologies had a better  
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28 agreement and lower variability. As more fuel efficient aircraft engines with low emission  
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30 combustors continue to be developed, instruments for measuring nvPM mass should have the  
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32 capability of higher resolution and sensitivity for low concentration levels. Future studies should  
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34 consider the variability associated with other instruments that meet the performance specifications  
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36 in AIR6241 but were not evaluated in this study.

37 It should be noted that the emission index values reported for nvPM mass and number have  
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39 not been corrected for size-dependent particles losses in the sampling and measurement systems,  
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41 and hence do not represent the actual emissions at the engine exit plane. Including a traceable size  
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43 measurement in the standardized measurement system would enable a more accurate estimation  
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45 of engine exit plane nvPM emissions to improve airport emissions inventory development and  
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47 environmental impact assessment of aircraft engine nvPM emissions. Size distribution  
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49 measurements, not currently specified in the standard method, were found to be approximated to  
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51 lognormal distributions with GMD ranging 15nm - 38nm, and GSD varying 1.53 - 1.92.

52 The wide range of ambient conditions encountered during the campaign affected the nvPM  
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54 emissions produced by the engine. A correction for changes in ambient conditions will need to be  
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56 developed to decouple the variability in the ambient temperature from the measured nvPM mass  
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58 and number emissions. Although the CFM56-7B26/3 engine used in this study is the most widely  
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60 used engine in commercial aviation, other engine types could have different emissions profiles. It  
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4 is essential that the reference systems be compared using an aircraft engine source with a different  
5 emissions profile to validate the repeatability and intermediate precision of the sampling and  
6 measurement systems established in this study. Also, long term comparison of the reference  
7 systems should be undertaken since these systems will continue to be used to varying extents over  
8 time.  
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13 As a direct consequence of the results from this ~~project~~campaign, several changes to the  
14 instrument performance were implemented such as updating the resolution of the MSS instruments  
15 to 0.01  $\mu\text{g}/\text{m}^3$ , and lowering the limit of detection of the nvPM mass instruments from 3  $\mu\text{g}/\text{m}^3$  to  
16 1  $\mu\text{g}/\text{m}^3$ . The procedure to demonstrate the conformity of the nvPM mass instruments to  
17 performance specifications was updated to include an additional applicability criterion for  
18 validating the nvPM mass instrument calibration on aircraft turbine engine exhaust.  
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24 The results from this study are a benchmark for the variability in standardized sampling  
25 and measurement systems for measuring aircraft engine emissions. The three reference systems  
26 evaluated were subsequently used for comparisons with aircraft engine manufacturer sampling and  
27 measurement systems. The aircraft engine manufacturers contributed nvPM emissions datasets for  
28 24 aircraft engine types that were representative of the current commercial fleet for inclusion in a  
29 database (Agarwal et al., 2019). With the database and knowledge of the uncertainty as  
30 characterized by the intermediate precision, the new ICAO nvPM mass and number emissions  
31 regulatory standard for in production and new engines (CAEP/11) was developed. ~~Going forward,~~  
32 ~~this~~ This new ICAO regulatory standard will be used to certify all aviation engines with rated thrust  
33 > 26.7 kN for nvPM mass and number emissions performance.  
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